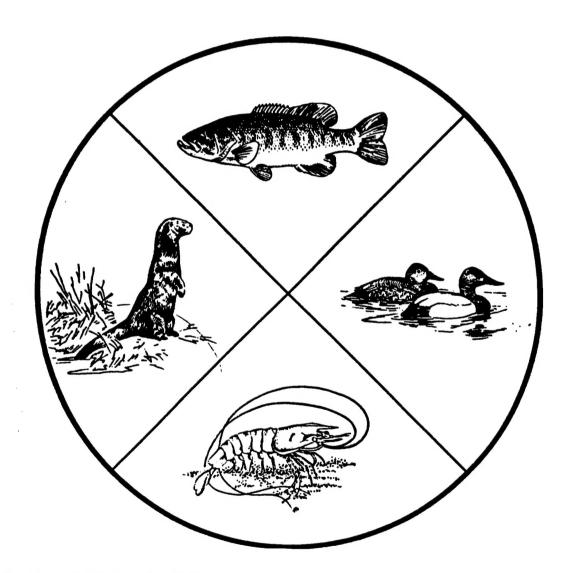
19970320 050

Acrolein Hazards to Fish, Wildlife, and Invertebrates: A Synoptic Review



National Biological Survey

U.S. Department of the Interior

DISTRIBUTION STATEMENT A

Approved for public release;
Distribution Unlimited

DTIC QUALITY INSPECTED 1

Technical Report Series

National Biological Survey

The National Biological Survey publishes five technical report series. Manuscripts are accepted from Survey employees or contractors, students and faculty associated with cooperative fish and wildlife research units, and other persons whose work is sponsored by the Survey. Manuscripts are received with the understanding that they are unpublished. Manuscripts receive anonymous peer review. The final decision to publish lies with the editor.

Editorial Staff

MANAGING EDITOR
Paul A. Opler

Assistant Branch Leader Paul A. Vohs

WILDLIFE EDITOR Elizabeth D. Rockwell

FISHERIES EDITOR
James R. Zuboy

VISUAL INFORMATION SPECIALIST Constance M. Lemos

> EDITORIAL CLERK Donna D. Tait

Series Descriptions

Biological Report

ISSN 0895-1926

Technical papers about applied research of limited scope. Subjects include new information arising from comprehensive studies, surveys and inventories, effects of land use on fish and wildlife, diseases of fish and wildlife, and developments in technology. Proceedings of technical conferences and symposia may be published in this series.

Fish and Wildlife Leaflet

ISSN 0899-451X

Summaries of technical information for readers of nontechnical or semitechnical material. Subjects include topics of current interest, results of inventories and surveys, management techniques, and descriptions of imported fish and wildlife and their diseases.

Fish and Wildlife Research

ISSN 1040-2411

Papers on experimental research, theoretical presentations, and interpretive literature reviews.

North American Fauna

ISSN 0078-1304

Monographs of long-term or basic research on faunal and floral life histories, distributions, population dynamics, and taxonomy and on community ecology.

Resource Publication

ISSN 0163-4801

Semitechnical and nonexperimental technical topics including surveys; data, status, and historical reports; handbooks; checklists; manuals; annotated bibliographies; and workshop papers.

Copies of this publication may be obtained from the Publications Unit, U.S. Fish and Wildlife Service, 1849 C Street, N.W., Mail Stop 130, Webb Building, Washington, DC 20240 (call 703-358-1711), or may be purchased from the National Technical Information Service (NTIS), 5285 Port Royal Road, Springfield, Virginia 22161 (call toll free 1-800-553-6847).

Acrolein Hazards to Fish, Wildlife, and Invertebrates: A Synoptic Review

 $\mathbf{B}\mathbf{y}$

Ronald Eisler

U.S. Department of the Interior National Biological Survey Washington, D.C. 20240

Contents

| Page |
|---|
| Abstract |
| Sources and Uses |
| General |
| Sources |
| Uses |
| Environmental Chemistry |
| General |
| Chemical Properties |
| Persistence |
| Metabolism |
| ethal and Sublethal Effects |
| General |
| Terrestrial Plants and Invertebrates |
| Aquatic Organisms |
| Birds |
| Mammals |
| Recommendations |
| acknowledgments |
| Cited Literature |
| Tables |
| |
| Tumber Page |
| 1 Chemical and other properties of acrolein |
| 2 Acrolein effects on representative aquatic organisms |
| 3 Acrolein effects on birds |
| 4 Acrolein effects on selected mammals |
| 5 Proposed acrolein criteria for the protection of living resources and human health 25 |

Acrolein Hazards to Fish, Wildlife, and Invertebrates: A Synoptic Review

by

Ronald Eisler

U.S. Department of the Interior National Biological Survey Patuxent Environmental Science Center Laurel, Maryland 20708

Acrolein (CH2=CHCHO) is the simplest member of the unsaturated aldehydes and enters the environment from incomplete combustion of fossil fuels, industrial discharges, herbicides, chemical control agents of fouling organisms, and normal metabolic processes of animals. Acrolein is volatile, flammable, and explosive. Biochemical and toxic effects of acrolein are caused by its reaction with sulfhydryl compounds to form a stable thiol ether. Acrolein metabolites under certain conditions are reportedly mutagenic, teratogenic, or carcinogenic. Acrolein degrades quickly in soils and in plant tissues; in water the half-time persistence is usually less than 50 h and in the atmosphere, less than 3 h. In treated irrigation canals, acrolein probably eliminates or seriously depletes all populations of aquatic fauna. Recommended herbicidal concentrations of acrolein for the control of submerged aquatic weeds usually exceed $1,000 \,\mu\text{g/L}$; however, short-term tests with various species show that frog tadpoles die at 7 μ g/L, representative fish are killed at 14 to 62 μ g/L, and sensitive crustaceans are immobilized or die at 34 to 80 µg/L. Terrestrial plants and insects are comparatively resistant to acrolein; terrestrial plants tolerated 500 μ g acrolein/L air and 25,000 μ g/L in irrigation water, and adult fruitflies (Drosophila melanogaster), 3,700,000 µg acrolein/L culture medium. Birds are adversely affected by concentrations greater than 51 µg acrolein/kg whole egg by injection of eggs, greater than $9{,}100\,\mu\text{g/kg}$ body weight (BW) by single oral doses, and greater than 50,000 µg/L (greater than 113 mg/m³) by air concentrations. Mammals were affected by $50\,\mu g$ acrolein/L air for 1 min, by $300\,\mu g/L$ air for 10 min, and by intravenous injections of 850-6,000 µg/kg BW. Acrolein was fatal to mammals after exposure to $660 \,\mu\text{g/L}$ air for 24 days, $8,000-11,000 \,\mu\text{g/L}$ air for 4 h, 875,000μg/L air for 1 min, and 4,000-28,000 μg/kg BW by single oral doses, or when fed diets equivalent to 500 µg/kg BW for 102 weeks. Proposed acrolein criteria for the protection of various resources include less than 15,000 µg/L in irrigation water of agricultural crops, less than 68 μ g/L for aquatic fauna in acute exposures and less than 21 μ g/L in chronic exposures, and less than $44 \, \mu \text{g/L}$ (less than $0.1 \, \text{mg/m}^3$) in air for rats. No acrolein criteria are now available for the protection of avian and terrestrial wildlife. Acrolein criteria for the protection of human health include less than 320 ug/L in drinking water, less than 110 μ g/L in air (less than 0.25 mg/m³), and less than 0.68 μ g/kg BW daily intake from all sources. More research is needed on acrolein and its metabolites.

Key words: Acrolein, aldehyde, herbicide, ecotoxicology, fish, aquatic plants, invertebrates, criteria.

Acrolein (CH₂=CHCHO) is an aldehyde that was first isolated in 1843 from the dry distillation of fats and glycerol (Beauchamp et al. 1985). It is now known that acrolein is ubiquitous in the environ-

ment; it is often present in trace amounts in foods and as a component of smog, fuel combustion products such as wood smoke, exhaust emissions from internal combustion engines, and cigarette smoke (Smith 1962; U.S. Environmental Protection Agency [EPA] 1980; Beauchamp et al. 1985). Atmospheric concentrations of acrolein over urban areas are between 2 and 7 ug/L: cigarette smoke, however contains about 10,000 µg of acrolein/L (Beauchamp et al. 1985). Acrolein is classified as a hazardous chemical because of its reactivity and flammability (EPA 1980). At low sublethal concentrations, acrolein is widely known for its acrid pungent odor and strong irritating effects on mucous membranes of the eyes and of the upper respiratory tract, its toxicity to cilia in all organisms, and its interference with nucleic acid synthesis in bacteria (Marano and Puiseux-Dao 1982; Beauchamp et al. 1985). In bulk, acrolein during storage or transfer is potentially hazardous if it becomes overheated or contaminated with water. For example, in 1982, 17,000 residents from Toft, Louisiana, were evacuated when two large tanks of acrolein began to burn (Bowmer and Smith 1984).

Acrolein enters the aquatic environment from its use as an aquatic herbicide, from industrial discharges, and as a byproduct of the chlorination of organic compounds in wastewater and drinking water treatment (EPA 1980). Dilute solutions of acrolein kill undesirable plant life in irrigation streams and ditches (National Research Council [NRC] 1977) and have been used routinely in about 4,000 km of irrigation canals in southeastern Australia to control submerged weeds, including Potamogeton tricarinatus, Elodea canadensis, and Vallisneria gigantia (Bowmer and Smith 1984). Acrolein has also been used for many years in channel maintenance in the United States (especially in western states), Canada, Egypt, Argentina, Mexico, and Turkey (Bowmer and Smith 1984). Unlike most other aquatic herbicides, acrolein rapidly dissipates from the water by volatilization and degradation without leaving phytotoxic residues (Bowmer and Smith 1984; Parent et al. 1992). However, acrolein provides only temporary control of submerged weeds and also kills fish and other aquatic life at recommended treatment concentrations (Bowmer and Smith 1984). In one Montana stream, acrolein killed all fish in a 4-km stretch after application to control submerged weeds; some fish deaths were recorded as far as 6.4 km downstream (Fritz-Sheridan 1982). Useful reviews on ecological and toxicological aspects of acrolein are presented by Smith (1962), EPA (1980), Beauchamp et al. (1985), and the Agency for Toxic Substances and Disease Registry [ATSDR] (1990).

This report is part of a continuing series of brief reviews of environmental contaminants and their effects on living organisms with special emphasis on fishery and wildlife resources. It was prepared in response to requests for information on acrolein from environmental specialists of the U.S. Fish and Wildlife Service.

Sources and Uses

General

Acrolein enters the environment as a result of normal metabolic processes; incomplete combustion of coal, wood, plastics, tobacco, and oil fuels; and industrial emissions. Acrolein has been detected in smog, foods, and water. It is used extensively in chemical manufacture, for control of fouling organisms, and as an herbicide to control submerged weeds in irrigation canals.

Sources

Acrolein is ubiquitous in the environment as a result of natural and anthropogenic sources. Sources of atmospheric acrolein include smog; incomplete combustion of coal, wood, gasoline, plastics and fats; tobacco smoke; and industrial emissions. The total amount of acrolein released into the atmosphere is unknown. In 1978, production losses of acrolein by emission from the four main U.S. plant locations were estimated at 34,682 kg; however, the gaseous emission streams are now either burned on emergence from the exhaust stack or sent to a furnace to destroy residual material (Beauchamp et al. 1985). Acrolein is found in photochemical smog and contributes to the smog's irritant capacity to the eye and respiratory pathways (Beauchamp et al. 1985; Leikauf et al. 1989). Recorded maximum acrolein concentrations in smog ranged from 12 to $14 \,\mu g/L \,(0.025 \,\text{to}\, 0.032 \,\text{mg/m}^3)$ in Los Angeles between 1961 and 1963 and were 13 µg/L in Hudson County, New Jersey (EPA 1980). For humans, exposure to atmospheric acrolein is greatest in the vicinity of incompletely combusted organic materials such as coal, wood, and petrol; highest acrolein concentrations are reported near forest fires and urban area fires (Beauchamp et al. 1985: Srivastava et al. 1992). The burning of southern pine (Pinus sp.), for example, generates 22 to 121 mg of acrolein/kg of burned wood (EPA 1980). Acrolein is also in the smoke of burning plastic materials. Air samples from more than 200 fires in Boston, Massachusetts, contained greater than

3,000 µg acrolein/L (greater than 6.8 mg/m³) in more than 10% of all samples; greater than 3,000 µg acrolein/L air is an immediately hazardous concentration for human life and health (Beauchamp et al. 1985). Cigarette smoke in some enclosed areas may account for as much as 12,400 µg of acrolein/L air (Feron et al. 1978; Astry and Jakab 1983; Beauchamp et al. 1985; Leikauf et al. 1989; Cohen et al. 1992). In the case of an enclosed room of 30 m³ capacity, smoking 5 cigarettes raises the air concentration to about 50 ug acrolein/L and smoking 30 cigarettes, to 380 µg/L (EPA 1980).

Acrolein is also generated when animal or vegetable fats are subjected to high temperatures (Feron et al. 1978; EPA 1980). Acrolein was detected aboard submarines in trace concentrations as a degradation product during the heating of lubrication oils and edible fats (Lyon et al. 1970). Large amounts of acrolein are generated from exhausts of internal combustion engines (Astry and Jakab 1983; Heck et al. 1986; Ballantyne et al. 1989). Acrolein concentrations of 10,000 µg/L (23 mg/m³) have been measured in nondiesel automobile exhausts, 2,900 μ g/L in diesel engine emissions, and 2,600-9,600 µg/L in other internal combustion engines (EPA 1980). Acrolein concentrations in air from several urban areas in the United States ranged from a maximum of 10 µg/L in 1960 to 1.8-3.4 µg/L in 1968; air in Tokyo during this period had an average acrolein concentration of 7.2 µg/L (Beauchamp et al. 1985). Urban acrolein pollution is derived mainly from automobile exhaust and incomplete burning of refuse (Beauchamp et al. 1985). Acrolein is formed during normal metabolic degradation of spermine and spermidine, glycerol, allyl formate, allyl alcohol, and cyclophosphamide (EPA 1980; Marano and Puiseux-Dao 1982; Leach et al. 1987). Acrolein was also in spores from the wheat stem fungus (Puccinia graminis) of infected wheat (Triticum aestivum); acrolein was the major chemical factor that normally induced infection structure formation in Puccinia (Macko et al. 1978).

Acrolein has been detected in effluent-water streams from industrial and municipal sources. Municipal effluents from Dayton, Ohio, for example, contained between 20 and 200 µg acrolein/L in 6 of 11 analyzed samples (EPA 1980; Beauchamp et al. 1985). Acrolein is also a component of many foods, and processing may increase the acrolein content (EPA 1980). Acrolein has been identified in raw turkey, potatoes, onions, coffee grounds, raw cocoa beans, alcoholic beverages, hops (EPA 1980), white bread, sugarcane molasses, souring salted pork, and cooked bluefin tuna (Thunnus thynnus; Beauchamp et al. 1985).

Occupational exposure to acrolein may occur during its production and isolation as a chemical intermediate or during the manufacture of acrylic acid, acrylic acid esters, and methionine (Beauchamp et al. 1985). Other sources of acrolein in the workplace include emissions from rubber vulcanization plants, from welding of metals treated with anticorrosion primers, and from pitchcooking plants and skin contact with herbicides during applications for aquatic weed control and with paper and paperboard, the manufacture of which includes acrolein as a slimicide. Acute acrolein poisoning from occupational exposure is improbable. However, the risks of poisoning are significant in certain industries including welding of fat and oil cauldrons, smelting work and foundry operations, printing plants, linoleum and oil cloth factories, manufacture of insulators, tin plating of sheet iron, and processing of linseed oil (Beauchamp et al. 1985).

Uses

Since its discovery in 1843, acrolein has been known to polymerize readily in the presence of many chemicals, and since 1947 it has been used safely in a wide variety of commercial applications (Albin 1962; Fischer 1962). Acrolein is presently produced by the catalytic oxidation of propylene for the manufacture of methionine, glutaraldehyde, 1,2,6-hexane thiol, and other chemicals. The largest quantity of acrolein produced by this process is converted directly to acrylic acid and acrylic acid esters (Beauchamp et al. 1985). In 1975, global production of acrolein was 59,000 metric tons; in 1980, this value was 102,000 tons-including 47,600 tons produced by the United States (EPA 1980). In 1983, about 250,000 tons (about 550 million pounds) of acrolein were produced; 92% was converted to acrylic acid and 5% to methionine; 3% was used as an aquatic herbicide (Beauchamp et al. 1985; Heck et al. 1986). Acrolein copolymers are used in photography, in textile treatment, in the paper industry, as builders in laundry and dishwater detergents, and as coatings for aluminum and steel panels (EPA 1980). Acrolein is used to scavenge sulfides from oil-field floodwater systems (Kissel et al. 1981), to crosslink protein collagen in the leather tanning industry, and to fixate tissue of histological samples (EPA 1980). The use of acrolein as a military poison gas has been advocated because

Acrolein has been used since 1960 to control submerged aquatic weeds in irrigation systems in the United States, Australia, and other countries where open channels distribute water for crop production (Hill 1960; Bartley and Hattrup 1975; Bowmer and Higgins 1976; EPA 1980; Reinert and Rodgers 1987). Acrolein is preferable to sodium arsenite for herbicidal control of submerged weeds because arsenicals are persistent (as long as for 1 year) and the high arsenic concentrations that are attained in water may be hazardous to humans and livestock (Hill 1960). Acrolein is extremely effective in killing submerged weeds that are difficult to control with other herbicides (Hill 1960). Acrolein has also been used as an herbicide in ponds, drains, and other bodies of water (Donohue et al. 1966). In Australia, the concentration of acrolein in irrigation canals to control various species of Elodea, Potamogeton, and Vallisneria is usually less than 15,000 µg/L (Bowmer and Higgins 1976). In general, acrolein has a low order of toxicity to terrestrial plants (Donohue et al. 1966). Most field and garden crops can tolerate water with as much as 15,000 µg acrolein/L without serious adverse effects (Bartley and Hattrup 1975). Acrolein, as discussed later, has comparatively low persistence and low accumulation in aquatic ecosystems. One disadvantage to its use as an herbicide is its pungent, irritating odor (Hill 1960). At recommended treatment concentrations, however, acrolein kills fish and other aquatic organisms; therefore, acrolein should be used only in aquatic systems where these resources are considered expendable (Reinert and Rodgers 1987).

Acrolein has been used to control bacteria, fungi, algae, and molluscs in cooling-water systems: 1,500 µg/L killed as much as 95% of the target species in a once-through treatment (Donohue et al. 1966). Acrolein has been applied directly to the marine environment to control the growth and settlement of mussels (*Mytilus edulis*) and other fouling organisms in cooling-water systems of coastal steam-electric-station power plants (EPA 1980; Rijstenbil and van Galen 1981). Mussels in marine cooling-water systems are controlled with 600 µg acrolein/L for 8 h daily for 3 days or with 700 µg/L for 3 h daily for 2 weeks (Rijstenbil and van Galen 1981). Acrolein prevents the growth

of microorganisms in liquid fuels such as jet fuels, in feed lines of subsurface wastewater injectors, and in water conduits of paper manufacturing plants (EPA 1980; Beauchamp et al. 1985).

Environmental Chemistry

General

Acrolein, the simplest member of the class of unsaturated aldehydes, has a pungent, irritating odor. It is volatile, flammable, and explosive and requires elaborate and specific conditions for storage and use. The half-time persistence of acrolein in freshwater is usually less than 50 h; in seawater it is less than 20 h and in the atmosphere less than 3 h. Biochemical and toxic effects of acrolein are caused by its rapid and essentially irreversible reaction with sulfhydryl compounds to form a stable thiol ether; however, many compounds can mitigate or block its toxicity. Acrolein is eventually metabolized to acrylic acid and glyceraldehyde; glycidaldehyde—an intermediate metabolite with mutagenic and carcinogenic properties—has been produced in vitro but not in vivo.

Chemical Properties

Acrolein is soluble in water and in many organic solvents including ethanol, acetone, and ether (Table 1; Beauchamp et al. 1985). Acrolein is a highly reactive molecule with two reactive centers: one at the carbon-carbon double bond and the other at the aldehydic group. Typical reactions involving acrolein are shown in detailed figures in Beauchamp et al. (1985). Acrolein is extremely volatile, flammable, and explosive (Table 1; Reinert and Rodgers 1987), especially in sunlight or in the presence of alkali or strong acid (Albin 1962; EPA 1980). A potential hazard in handling acrolein is its rapid exothermic polymerization caused by the use of insufficient hydroquinone inhibitor or lack of strict control of pH (Beauchamp et al. 1985). Commercial acrolein should be maintained at pH 6.0 and contain less than 3% water and 0.1-0.25% hydroguinone as a polymerization inhibitor. Atypical commercial sample contains about 97% acrolein, 0.5% other carbonyls, and 2.5% water. The addition of hydroquinone (0.1-0.25%) prevents the vinyl polymerization of acrolein, and controlling the pH between 5 and 6 by acetic acid increases stability of commercial acrolein by preventing aldol condensation. Elaborate and specific conditions are now prescribed for the storage of

Table 1. Chemical and other properties of acrolein a

| Variable | Data |
|--|---|
| Chemical name | 2-Propenal |
| Alternate names | Acraaldehyde, acraldehyde, acrolein, acryladehyde, acrylaldehyde, acrylic aldehyde, allyl aldehyde, aqualin, aquilin, Magnacide H, propenal |
| CAS Number | 107-02-8 |
| Structural formula | CH_2 = $CHCHO$ |
| Molecular weight | 56.06 |
| Specific gravity | 0.8427-0.8442 |
| Physical state | Colorless or yellow liquid at 25° C |
| Odor | Pungent, irritating |
| Boiling point | 52.5–53.5° C |
| Melting point | −86.95° C |
| Solubility | |
| Water | 206–208 g/L |
| Organic solvents | Miscible |
| $\operatorname{Log} K_{\operatorname{ow}}$ | 0.01 |
| Vapor pressure | 215–220 mm Hg at 20° C |
| Explosive limits of | |
| vapor and air | |
| Upper limit | 31% acrolein |
| Lower limit | 2.8% acrolein |

^a Hill (1960), Anderson and Hood (1962), Folmar (1977), EPA (1980), Hudson et al. (1984), Beauchamp et al. (1985), Mayer (1987), Reinert and Rodgers (1987), Ballantyne et al. (1989), Sine (1991), Agency for Toxic Substances and Disease Registry [ATSDR] (1990), National Institute for Safety and Health [NIOSH] (1990).

acrolein and include vents and safety valves, construction materials, fire control, spills, and waste disposal (Beauchamp et al. 1985). Commercial acrolein is stored and shipped under a blanket of oxygen-free inert gas (Albin 1962).

Spectrophotometric determination with 4hexylresorcinol and a fluorometric method with m-aminophenol are the most commonly used procedures for the determination of acrolein; however, gas chromatography and high performance liquid chromatography procedures are also used (EPA 1980; Kissel et al. 1981; Nishikawa and Hayakawa 1986). Acrolein concentrations in rainwater between 4 and 200 µg/L can be measured rapidly (in less than 80 min) without interference from related compounds; the method involves acrolein bromination and analysis by gas chromatography with electron capture detection (Nishikawa and Hayakawa 1986). Kissel et al. (1981) emphasize that water samples from potential acrolein treatment systems require the use of water from that system in preparing blanks, controls, and standards and that acrolein measurements should be made at the anticipated use concentrations.

Persistence

Degradation and evaporation seem to be the major pathways for acrolein loss in water; smaller amounts are lost through absorption and uptake by aquatic organisms and sediments (EPA 1980; Reinert and Rodgers 1987). The half-time persistence of acrolein in freshwater is 38 h at pH 8.6 and 50 h at pH 6.6; degradation is more rapid when initial acrolein concentrations are less than 3,000 µg/L (Bowmer and Higgins 1976). At pH 5, acrolein reacts by reversible hydrolysis to produce an equilibrium mixture with 92% beta-hydroxypropionaldehyde and 8% acrolein; in alkali, the primary reaction is consistent with a polycondensation reaction (EPA 1980). In natural waters, acrolein degradation proceeds to carboxylic acid via a microbial pathway (EPA 1980); beta-hydropropionaldehyde is readily biotransformed in about 17.4 days (Reinert and Rodgers 1987).

Acrolein is applied to irrigation canals to control submerged aquatic weeds at greatly different time-concentration treatments. Regardless of time-concentration regimens—which vary from 100 µg/L for 48 h in the United States to 15.000 µg/L for several hours in Australia—the daily decay rate constants are remarkably similar, ranging from 0.14 to 0.21, and are probably affected by variations in weed density (O'Loughlin and Bowmer 1975; Parent et al. 1992). In one case, acrolein applied to the Columbia River at an average initial concentration of 125 µg/L degraded to 25 µg/L after 48 h in samples greater than 65 km from the application point—a loss of 80% (EPA) 1980). High initial concentrations (50,000-160,000 µg/L) of acrolein in natural waters degraded 57 to 80% in 192 h, suggesting that high concentrations can alter the rate of hydrolysis (Kissel et al. 1981). In seawater, the half-time persistence of acrolein was less than 20 h (Rijstenbil and van Galen 1981). In photochemical smog, acrolein is comparatively unstable and not likely to persist; the dominant removal mechanism involves hydroxide attack on acrolein, and the atmospheric half-life persistence is 2-3 h under these conditions (Beauchamp et al. 1985).

Metabolism

Biochemical and toxic effects of acrolein are probably caused by its reaction with critical protein and nonprotein sulfhydryl groups (EPA 1980; Beauchamp et al. 1985; Heck et al. 1986). The reaction of acrolein with sulfhydryl compounds is rapid and essentially irreversible, resulting in the formation of a stable thiol ether (Beauchamp et al. 1985: Heck et al. 1986). Metabolism of acrolein is believed to result in the formation of acrylic acid and glyceraldehyde (Figure). The postulated metabolites of acrolein can be oxidized to carbon dioxide (Beauchamp et al. 1985). Acrylic acid does not seem to represent a significant toxic hazard when compared with the parent acrolein because at low airborne concentrations of less than 1,000 µg acrolein/L, the quantity of acrylic acid produced by metabolism is negligible. Thus, metabolism to acrylic acid after inhalation should be regarded as a detoxification pathway. Conjugation of acrylic acid with glutathione represents another elimination and detoxification pathway (Beauchamp et al. 1985). In-vitro studies of acrolein metabolism in mammals suggested that acrolein exposures may result in exposure to glycidaldehyde, an intermediate in acrolein metabolism (Figure). The major toxic effects of acrolein exposure-including irritation, ciliastasis, and hypersensitivity—are probably due either to the parent acrolein or to the reaction of glycidaldehyde with cell proteins. Glycidaldehyde is a potent mutagen and carcinogen; however, no evidence is available showing that acrolein can produce glycidaldehyde in vivo (Beauchamp et al. 1985). Acrolein is more toxic when inhaled than when taken orally (EPA 1980). Inhalation of acrolein decreased the concentrations of protein and nonprotein sulfhydryl groups in nasal mucosal tissue (Heck et al. 1986). Acrolein is highly reactive towards thiol groups and rapidly conjugates with glutathione and cysteine (EPA 1980). When glutathione is depleted, acrolein potentiates the nasal toxicity of formaldehyde to rats (Heck et al. 1986).

Acrolein is a metabolite of allyl alcohol and cyclophosphamide, and these compounds should be considered in acrolein metabolism schemes (Beauchamp et al. 1985; Cohen et al. 1992). Allyl alcohol in the presence of NADPH and liver or lung microsomes degrades to acrolein, acrylic acid, and glycidol (Figure).

When added to water as an aquatic herbicide, acrolein undergoes rapid decomposition, especially in the sunlight. At the same time, it reacts

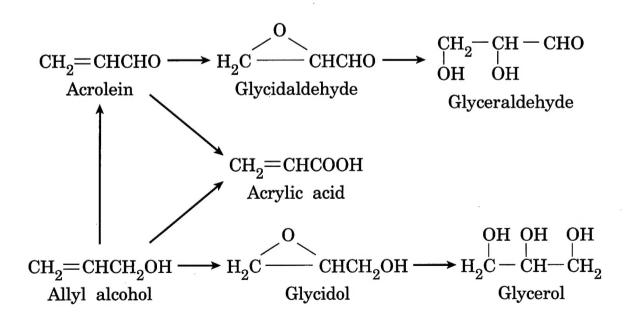


Figure. Proposed scheme for in vitro mammalian metabolism of acrolein and allyl alcohol, a precursor of acrolein (Beauchamp et al. 1985; ATSDR 1990).

rapidly with amines, alcohols, and mercaptans of aquatic plants, destroying cell structure and killing the plants (Parent et al. 1992). Mammals drinking acrolein-contaminated water rapidly convert acrolein to saturated alcohol compounds because of the low pH in the upper portion of their GI tracts; the primary breakdown product is betapropionaldehyde (EPA 1980).

Many compounds including glutathione, 2-mercaptoethanol, beta-dimethylcysteamine, penicillamide, gamma-mercaptopropionylglycine, and N-acetylcysteine mitigate or block the toxic effects of acrolein (EPA 1980; Beauchamp et al. 1985; Heck et al. 1986). In frogs (Rana japonica), sulfhydryl compounds reduce the effects of acrolein on excitation-contraction uncoupling in skeletal muscle (Fujino et al. 1985). In mice, cysteine reduced the cytotoxic effects of acrolein on tumor cells; in rabbits, cysteine mitigated acrolein-induced alveolar macrophage calcium-dependent ATP-ase, phagocytosis, and adhesiveness (EPA 1980). In male rats, cysteine and ascorbic acid antagonized the acute lethal effects of orally-administered acrolein, and 2-mercaptoethanol antagonized the inhibitory effect of acrolein on liver DNA-polymerase (EPA 1980).

Lethal and Sublethal Effects

General

Acrolein degrades quickly in soils and in plant tissues regardless of mode of administration. Most terrestrial crop plants easily tolerate 25,000 µg of acrolein/L of irrigation water and some can tolerate 70,000-80,000 µg/L without adverse effects. Terrestrial plants were adversely affected at atmospheric concentrations of 500 µg acrolein/L air, but this concentration exceeds the recommended value of 110 μg/L (0.25 mg/m³) air for protection of human health in occupational settings.

Adult fruitflies (Drosophila melanogaster) were comparatively resistant to acrolein and had lowered survival when reared in culture media with greater than 3,700,000 µg acrolein/L. At recommended concentrations for the control of nuisance submerged aquatic weeds (frequently 100-1,000 μg/L, often greater than 9,600 μg/L), acrolein was lethal or harmful to almost all aquatic vertebrates and invertebrates tested in short-term exposures. The most sensitive groups of tested aquatic organisms in short-term assays were frog tadpoles (dead at 7 µg/L) and representative species of fish

(reduced survival at 14-62 µg/L) and crustaceans (death or immobilization at 34-80 µg/L), Adverse effects of acrolein on birds were observed at acute oral doses of 9,100 µg/kg BW (reduced survival), at concentrations greater than 51 µg/kg egg for egg injection (abnormal development and reduced survival), and at greater than 50,000 ug/L air (respiratory tract histopathology). In mammals, acrolein is a strong cytotoxic and ciliostatic agent that is irritating to mucous membranes of dermal, ocular, gastrointestinal, and respiratory systems and is systemically toxic by all routes of exposure. Adverse effects of acrolein are documented in sensitive species of mammals under the following regimens: 50 µg/L air for 1 min (increased blood pressure and heart rate); 300 µg/L air for 10 min (ocular and nasal irritation); 500 to 1,000 µg/L air (repelled by odor); 660 µg/L air for 24 days (reduced survival); 8,000 to 11,000 µg/L air for 4-6 h or 875,000 µg/L air for 1 min (death); dietary concentrations equivalent to 500 µg/kg BW for 102 weeks (decreased survival); 850-6,000 µg/kg BW by intravenous injection (liver necrosis, embryo resorption); and single oral doses between 4,000 and 28,000 µg/kg BW (death).

Acrolein was mutagenic to certain microorganisms and to the fruitfly; mutagenicity may be due. in part, to glycidaldehyde, an acrolein metabolite. Injected into the amniotic fluid, acrolein is teratogenic to rats; teratogenicity may be due to acrylic acid, an acrolein metabolite. There is limited evidence that acrolein acts as a weak carcinogen and tumor promoter. Acrolein interacts with other chemicals, sometimes synergistically, additively, or antagonistically. Also, some chemicals normally contain acrolein as an impurity or yield acrolein as a metabolite.

Terrestrial Plants and Invertebrates

Most crop plants easily tolerate irrigation water with 25,000 µg of acrolein/L and many tolerate 70,000 to 80,000 µg/L without adverse effects—including corn (Zea mays), cotton (Gossypium hirsutum), milo (Sorghum spp.), squash (Cucurbita spp.), castor bean (Ricinus communis), tomato (Lycopersicon esculentum), alfalfa (Medicago sativa), and sugarcane (Saccharum officinarum; Ferguson et al. 1961). Acrolein degrades quickly in soils and plant tissues regardless of mode of administration (Ferguson et al. 1961). Atmospheric concentrations of 500 µg acrolein/L and higher were harmful to certain plants (Beauchamp et al. 1985). Leaves of the pinto bean (*Phaseolus* spp.) and the morning glory (*Ipomoea* spp.) developed brown foliar lesions after exposure to 500 μ g/L air for 4–7 h; damage was more severe if the plants were moist during exposure. Leaves of the radish (*Raphanus* spp.) developed lesions after exposure to 1,500 μ g acrolein/L air for 6–7 h; however, leaves of the geranium (*Geranium* spp.) and the tomato showed no adverse effects after exposure to 1,500 μ g/L air for 7 h (Beauchamp et al. 1985).

Acrolein inhibits DNA, RNA, and protein synthesis in the bacterium Escherichia coli, and this inhibition probably accounts for its cytotoxic and inhibitory effects on E. coli cell division (EPA 1980; Beauchamp et al. 1985). Acrolein is demonstrably mutagenic to microorganisms and to larvae of the fruitfly (Drosophila melanogaster). Acrolein-induced mutagenicity-including point mutations, sister chromatid exchanges, and chromosome breakages—has been observed in selected strains of bacteria (E. coli, Salmonella typhimurium), yeast (Saccharomcyces cerevisiae), fruitfly larvae, and cultured Chinese hamster ovary cells (EPA 1980; Beauchamp et al. 1985; Sierra et al. 1991; Cohen et al. 1992). Acrolein's mutagenicity may be due to the metabolite glycidaldehyde; glycidaldehyde was mutagenic to bacteria and yeast under controlled conditions (Beauchamp et al. 1985; Sierra et al. 1991). Studies with D. melanogaster show that acrolein is mutagenic in the sex-linked recessive lethal test when injected but not when fed (Sierra et al. 1991). Acrolein caused 2.2% sexlinked mutations in D. melanogaster—the highest percentage recorded among several tested aldehydes (EPA 1980). In studies by Comendador (1984), early embryonic stages of fruitflies were most sensitive to the mutagenic properties of acrolein and sensitivity decreased with increasing development to the point that adults showed negligible mutagenic responses. Adults of the fruitfly were generally resistant to acrolein; mortality was 25% when the culture medium contained $3.700.000 \, \mu g$ of acrolein/L, 50% at $8,600,000 \, \mu g/L$, and 75% at 22,100,000 µg/L (Comendador 1984).

Aquatic Organisms

Adverse effects of acrolein on sensitive groups of aquatic organisms are documented (Table 2) at concentrations—in µg acrolein/L medium—as low as 7 for frog tadpoles (death), 14–62 for fish (death), 34–80 for crustaceans (death, immobilization), 50 for oysters (reduction in shell growth rate), 100–200 for freshwater algae (DNA and

RNA reduction, photosynthesis inhibition), 151 for gastropods (death), >151 for insects (death), 500-2,000 for macrophytes (leaf cell deterioration, death), 1,250 for trematodes (death of miracidia in 20 min), and 62,000 for bacteria (growth reduction). Aquatic vertebrates were more sensitive than invertebrates (Holcombe et al. 1987), and younger fish were more sensitive than older fish (Burdick et al. 1964). Aquatic insects do not avoid acrolein at concentrations that repel fish (Folmar 1978).

As an herbicide, acrolein is most effective in controlling dense accumulations of submerged weeds in habitats where waterflow is rapid and uniform, such as irrigation canals and rapidlyflowing streams (Ferguson et al. 1961). Acrolein is lethal to various genera of submerged plants (Hydrodictvon, Spirogyra, Potamogeton, Zannichellia, Cladophora, Ceratophyllum, Elodea, Chara, Najas) at 1,500 to 7,500 µg/L (Ferguson et al. 1961; Beauchamp et al. 1985). But some floating plants (Pistia, Eichornia, Jussiaea) are more resistant to acrolein than submerged plants and require concentrations that are at least double those necessary for submerged forms (Ferguson et al. 1961). Acrolein has little effect on emergent aquatic macrophytes and should not be used to control emergents (Ferguson et al. 1961). In Australia, acrolein is the only herbicide now used for control of submerged aquatic weeds in larger irrigation canals (Bowmer et al. 1979); effective plant control was obtained at 9.6-28.8 mg/L for 3 h (Bowmer and Smith 1984). In the United States, the U.S. Bureau of Reclamation controls aquatic algae and weeds at lower concentrations (0.1 mg/L) and longer exposures (48 h; Folmar 1980). In the Columbia River basin in the state of Washington, acrolein is used to control submerged aquatic macrophytes at concentrations of 0.1 mg/L for 48 h or 1.0 mg/L for 4 to 8 h with applications every 3 to 5 weeks (Bartley and Hattrup 1975). Vegetation destruction by acrolein is maximal 1 week after application, and green filmenentous algae are usually the first plants to return after 1 month (Ferguson et al. 1961). Biomass and species diversity were altered in acrolein-treated phytoplankton populations in Egyptian irrigation canals 1 year after treatment (Kobbia 1982). Although acrolein is a powerful cytotoxic agent, its inhibitory effects at sublethal concentrations on plant mitosis, nucleic acid synthesis, and protein synthesis are considered completely reversible (Marano and Puiseux-Dao 1982).

Table 2. Acrolein effects on representative aquatic organisms.

| Caxonomic group, organism, concentration, and other variables | Effect | Reference ^a |
|---|--|------------------------|
| Bacteria, Algae, and Macrophytes | | |
| Fresh water algae, <i>Anabaena</i> sp.; 690 µg/L; 24-h exposure | 50% reduction in photosynthesis at $25^{\circ}\mathrm{C}$ | 1 |
| Aquatic bacteria, 3 species | | |
| $62,000 \mu g/L; 48-h exposure$ | Some growth reduction, but recovery by 120 h | 2 |
| 125,000 μg/L; 120-h exposure 500,000 μg/L; 2-h exposure | LC100 LC100 | 2 2 |
| Freshwater alga, Cladophora glomerata | | |
| 100 μg/L | Onset of photosynthesis inhibition at 30° C | 1 |
| 760 μg/L; 24-h exposure | 50% reduction in photosynthesis at 30° C | 1 |
| 1,000 µg/L; 24-h exposure | 50% reduction in photosynthesis at 25° C | . 1 |
| Alga, Dunaliella bioculata | - | |
| 100 μg/L; 48-h exposure | DNA concentration reduced 28% | 3 |
| 200 μg/L; 48-h exposure | DNA concentration reduced 36% and RNA 28% | 3 |
| 400 μg/L; 48-h exposure | DNA reduced 93%, RNA 68%, and proteins 74% | 3 |
| 1,000 μg/L; 48-h exposure | No development in 48 h | 3 |
| 8,000 μg/L; 3-h exposure | Ultrastructural anomalies, and cytoplamic inclusions | 4 |
| Elodea, Elodea canadensis | | |
| Sublethal (actual exposure concentration and duration unknown) | Growth stimulation (from reduced competition by aufwuchs, bacteria, and epiphytic algae) | 5 |
| 500 μg/L; 24-h exposure | Leaf cell deterioration | 6,7 |
| 2,800 μg/L; 3-h exposure in irrigation canal | 80% reduction in density; recovery began in 17 days | 5 |
| 15,000 µg/L; 2–6 h exposure in Australian canals | Effective control for up to 21 km in flowing-water irrigation canals | 8 |
| 18,000 μg/l; 2 to 12 h exposure in smaller channels and up to 72 h in major canals; New South Wales | Effective control | 5 |
| 22,000 μg/L; 3-h exposure in | 94% reduction in | 5 |
| irrigation canal | biomass after 14 days | |
| Filamentous algae, unidentified | | |
| 500 μg/L; 5-months exposure; petroleum-refinery recirculating cooling-water system | Effective control | 9 |
| 1,000 µg/L; 20-h exposure in Arizona irrigation canal | Effective control for 2 weeks | 10 |
| 3,500 µg/L; 2-weeks exposure in petroleum refinery cooling water | Lethal | 9 |
| 5,000 μg/L; 1-week exposure in petroleum refinery cooling water | Lethal | 9 |

Table 2. Continued.

| axonomic group, organism, concentration, and other variables | Effect | Referencea |
|---|--|------------|
| Freshwater alga, Enteromorpha intestinalis | | |
| 1,800 μg/L; 24-h exposure | 50% reduction in photosynthesis at 25° C | 1 |
| 2,500 µg/L for 24 h | 50% reduction in photosynthesis at 20° C | 1 |
| >5,000 µg/L for 24 h | 50% reduction in photosynthesis at 15° C | 1 |
| Freshwater plants; 6 species of submerged plants, 2 species of floating plants, 4 groups of phytoplankton; irrigation drains, Egypt; 15,000–25,000 μg/L for 45 min, repeated 4 times; | Effective control of all plants within 2–7 days. Phytoplankton recovery over 1-year period was most rapid for the Cyanophycae, followed by the Bacilliarophyceae, Chlorophycae, and Euglenophyceae, and resulting in altered biodiversity when compared with a control canal | 11 |
| Submerged macrophytes, 3 species (Najas sp., Ceratophyllum sp., and Ipomoea sp.); $25{,}000~\mu\text{g/L}$ | All dead 1 week after application | 6 |
| Floating pondweed, Potamogeton carinatus | | |
| 2,000 μg/L for 12 h | LC50 | 12 |
| $10,000-15,000 \mu g/L \text{ for > 1 h (actual)}$ | LC50 | 12 |
| exposure time unknown) 15,000 µg/L for 1.7 h | LC50 | 12 |
| 22,000–26,000 μg/L | LC80 | 12 |
| for >1 h (actual exposure time unknown) | | |
| Pondweed, Potamogeton crispus; 20,000 µg/L for 5 h | All dead in 8 days | 6 |
| Pondweed, Potamogeton tricarinatus; 4,000 µg/L; 1-h exposure in irrigation canal | Minimum effective concentration | 5 |
| Submerged macrophyte, <i>Vallisneria gigantea</i> ; 26,000 µg/L for 1 h in irrigation canal | Minimum effective concentration | 5 |
| Ribbonweed, Vallisneria spiralis | | |
| 1,600 (95% confidence interval [CI] | 50% reduction in biomass | 12 |
| of 1,300–2,000) μg/L 3,700 (95% CI of 3,200–4,600) μg/L for 1 h | 80% reduction in biomass | 12 |
| nvertebrates | | |
| Snail, <i>Aplexa hypnorum</i> ; 151 µg/L for 96 h | Less than 50% mortality | 13 |
| Snail, Australorbis glabratus | | |
| 1,250 µg/L for 24 h | All adults and 90% of embryos survived | 7 |
| 2,500 μg/L for 24 h | 35% of adults and 40% of embryos died | 7 |
| 10,000 μg/L for 24 h | 98% of adults and 100% of embryos died | 6,7 |
| Barnacle, <i>Balanus ebarneus</i> ; 1,600–2,100 μg/L for 48 h | LC50 | 6 |
| American oyster, Crassostrea virginica; 50–55 µg/L for 96 h | 50% reduction in shell growth rate | 6, 14, 15 |

Table 2. Continued.

| Taxonomic group, organism, concentration, and other variables | Effect | Reference |
|--|---|-------------|
| Daphnid, Daphnia magna | | 44, |
| 17 – $34~\mu g/L$ | $MATC^{b}$ | 6, 16 |
| 51 (95% CI of 43 to 62) µg/L for 48 h | 50% immobilized | 13 |
| 57–80 μg/L for 48 h | LC50 | 6 |
| Mayfly, Ephemerella walkeri; 100 μg/L for 1 h | No avoidance of acrolein by nymphs | 17 |
| Freshwater snails, 3 species (Physa, Biomphalaria, Bulinus); 25,000 µg/L for 3.5–4 h | All dead | 14 |
| Common mussel, Mytilus edulis | | |
| 200-1,000 μg/L; exposed for as much as 8 h daily for 3 days | Valves closed immediately after start of exposure to acrolein regardless of concentration or duration; effect in 45% of mussels at 200 µg/L, 80% at 400 µg/L, and 90% at 600 µg/L | 8 |
| 600 µg/L; single 8-h exposure followed by 48-h of uncontaminated seawater | 70% of the mussels (1–2.5 mm shell length) in the cooling water systems of power plants became detached in 3 days vs.13% of controls | 18 |
| 600 μg/L; 8-h exposure daily for 3 days 600 μg/L; 29-h continuous exposure | 97% of mussels became detached 100% detachment | 18 18 |
| Brown shrimp, Penaeus aztecus | | |
| 100 μg/L for 48 h 100 μg/L for 48 h | LC50 50% loss in equilibrium | 14 6, 15 |
| Trematode, Schistosoma mansoni | | • |
| $1,250~\mu g/L$ for $20~min$ $2,500~\mu g/L$ | Killed all miracidia Killed all miracidia in 10 min, and all cercariae in 18 min | 7 7 |
| Midge, Tanytarsus dissimilis; 151 µg/L for 48 h | Less than 50% mortality | 13 |
| Vertebrates | | |
| Bowfin, <i>Amia calva</i> ; 62 µg/L for 24 h | LC50, fry | 14 |
| Goldfish, Carassius auratus | | |
| 80 μg/L for 24 h 1,000–2,000 μg/L for 3 h | LC50 Fatal | 6 14 |
| White sucker, Catostomus commersoni; 14 (95% CI of 8–25) µg/L for 96 h | LC50 | 13 |
| Longnose killifish, <i>Fundulus similis</i> ; 240 µg/L for 48 h | LC50 | 6,15 |
| Western mosquitofish, Gambusia affinis | | |
| 61 μg/L for 48 h | LC50 | 6, 14 |
| 149 μg/L for 24 h | LC50 LC50 | 6, 14 14 |

Table 2. Continued.

| Caxonomic group, organism, concentration, and other variables | Effect | Reference |
|---|--|-----------|
| Bluegill, Lepomis macrochirus | | |
| $13 \mu g/L$ for $28 days$ | Whole fish, bioconcentration factor of 344 | 6 |
| 33 (95% CI of 27-40) μg/L for 96 h | LC5 | 13 |
| 79 μg/L for 24 h | LC50 | 6, 19 |
| 90–100 μg/L for 96 h | LC50 | 6, 14 |
| Largemouth bass, Micropterus salmoides | | |
| 160 µg/L for 96 h | LC50 | 6, 14 |
| 183 µg/L for 24 h | LC50 | 14 |
| Rainbow trout, Oncorhynchus mykiss | | |
| 8 µg/L for 48 h | None dead | 20 |
| 16 (95% CI of 14–19) μg/L for 96 h | LC50 | 13 |
| 20, 50, or 100 μg/L; exposure for 4 h; trout collected 1, 4, and 7 days postexposure; cooked fillets evaluated for odor and taste by human panel | Unacceptable organoleptic qualities were recorded for fillets 1 and 4 days ($P = 0.05$) after treatment with 100 μ g/L; some unacceptable qualities were detected 1 and 4 days after treatment with 50 μ g/L, and at 7 days after treatment with 100 μ g/L | 17 |
| 29 (95% CI of 22-37) | LC50 | 21 |
| μg/L for 96 h | | |
| 48 μg/L for 48 h | LC32 | 6, 20 |
| 65 μg/L for 24 h | LC50, fingerlings | 6 |
| 77 μg/L for 20.5 h | LC50 | 21 |
| 90 μg/L for 4.8 h | No deaths | 20 |
| 96 μg/L for 48 h | All dead | 20 |
| 100 μg/L for 1 h | Avoidance by fry | 6, 14, 23 |
| 150 μg/L | Lethal | 22 |
| 240 μg/L for 4.8 h | LC10 | 20 |
| 410 µg/L for 4.8 h | LC70 | 20 20 |
| >500 µg/L for 4.8 h | All dead | |
| Chinook salmon, <i>Oncorhynchus</i> tshawytscha; 80 µg/L for 24 h | LC50 | 6, 19 |
| Fathead minnow, Pimephales promelas | | |
| 11.4–41.7 μg/L | $MATC^{b}$ | 6, 16 |
| 14 (95% CI of 8–25) μg/L for 96 h | LC50 | 13 |
| 84 μg/L for 6 days | LC50 | 6 |
| 115 µg/L for 48 h | LC50 | 6, 14 |
| 150 μg/L for 24 h | LC50 | 14 |
| Harlequin fish, <i>Rasbora heteromorpha</i> ; 130 μg/L for 48 h | LC50 | 14 |
| Brown trout, Salmo trutta | | |
| 46 μg/L for 24 h | LC50 | 6, 14, 19 |
| 1,500 µg/L for 76–138 min | All dead | 19 |
| 6,000 μg/l for 28–61 min | All dead | 19 |
| 16,000 μg/L for 15–39 min | All dead | 19 |
| Frog, Xenopus laevis, tadpoles; 7 (95% CI of 6–8) µg/L for 96 h | LC50 | 13 |

^a 1, Fritz-Sheridan 1982; 2, Starzecka 1975; 3, Marano and Puiseux-Dao 1982; 4, Baron-Marano and Izard 1968; 5, Bowmer and Smith 1984; 6, EPA 1980; 7, Ferguson et al. 1961; 8, Bowmer et al. 1979; 9, Donohue et al. 1966; 10, Corbus 1982; 11, Kobbia 1982; 12, Bowmer and Sainty 1977; 13, Holcombe et al. 1987; 14, Folmar 1977; 15, Mayer 1987; 16, Beauchamp et al. 1985; 17, Folmar 1978; 18; Rijstenbil and van Galen 1981; 19, Burdick et al. 1964; 20, Bartley and Hattrup 1975; 21; McKim et al. 1987; 22, Kissel et al. 1981; 23, Folmar 1976.

^bMaximum acceptable toxicant concentration. Lower value in each pair indicates highest concentration tested producing no measurable effect on growth, survival, reproduction, or metabolism during chronic exposure; higher value indicates lowest concentration tested producing a measurable effect.

Acrolein in concentrations sufficient to control nuisance submerged aquatic weeds may also kill snails, crayfish, shrimp, fish, and toads (Ferguson et al. 1961). In one case, acrolein was used to control Potamogeton and Chara in an Ohio farm pond during June 1960 (Hill 1960). Acrolein was applied at 16,100 µg/L to a 0.1 ha portion of the 0.7 ha pond. Within 1 h of application, many dead amphibian tadpoles and small bluegills (Lepomis macrochirus) were recovered. In 24 h, Chara had turned white and *Potamogeton* brown; both plant species seemed dead; fish were swimming in the treated area. In 72-96 h, several large dead walleyes (Stizostedium vitreum vitreum) were found. One week posttreatment, all algae and weeds in the treated area were dead; weeds were present in the untreated areas. The treated section remained weed-free for 4-6 weeks; after 8 weeks, the treated area was heavily infested with Chara. Hill (1960) concluded that tadpoles, walleyes, and small bluegills were more susceptible to acrolein toxicity than larger bluegills and bass (Micropterus spp.) in the pond. Acrolein is also effective in controlling trematodes that cause schistosomiasis wherein snails are the intermediate host, especially in irrigation systems. For example, native species of snails (Lymnaea, Helisoma), along with Potamogeton weeds, were destroyed within 12 km in the main irrigation canal of Kern County, California, after a single application of acrolein (Ferguson et al. 1961).

Acrolein was the most toxic of 15 herbicides tested for toxicity to fish (EPA 1980). Responses by rainbow trout (Oncorhynchus mykiss) surviving 77 µg acrolein/L, a concentration that killed 50% in about 21 h, were characteristic of respiratory irritants (McKim et al. 1987). These responses included a steady increase in cough rate; decreases in ventilation rate, oxygen utilization. and heart rate; increases in hematocrit; and decreases in total arterial oxygen, carbon dioxide, and pH (McKim et al. 1987). In studies by Bartley and Hattrup (1975), no-observable-effect concentrations of acrolein for rainbow trout were 240 μg/L for exposures of 4.8 h and 48 μg/L for exposures of 48 h; these values are below the concentrations that control aquatic weeds. In the same study, rainbow trout that survived exposure to high sublethal concentrations for 48 h were unable to recover completely after acrolein treatments were ended. Trout and other teleosts are poorly adapted to detoxify acrolein and other xenobiotic aldehydes (Parker et al. 1990). The low

metabolic capacity of fish liver aldehyde dehydrogenase for aldehydes, in general, suggests that these compounds may be hazardous to fish populations (Parker et al. 1990). Applications of acrolein to waters where fish may be taken for human consumption should be made with caution; rainbow trout surviving exposure to acrolein in reservoirs or connecting canals frequently presented odor and taste problems to human consumers (Folmar 1980).

Acrolein is used also to control fouling organisms in cooling water systems. Effective control was established in a once-through cooling system of a steel mill with continuous application of 200 µg acrolein/L (Donohue et al. 1966). Acrolein controlled bacteria in condenser pipes of a powerplant cooling system but only at extremely high concentrations of 125,000 µg/L for 120 h or 500,000 µg/L for 2 h (Starzecka 1975). Acrolein reduced settlement of young mussels (*Mytilis* sp.) in cooling seawater systems of power plants (Rijstenbil and van Galen 1981). In recirculating cooling water systems, algae and bacteria can be controlled at 500 µg/L for 5 months or at 5.000 µg/L for 1 week (Table 2).

Birds

Acrolein was lethal to birds at single oral doses of 9,100 µg/kg BW (Table 3). Observed signs of acrolein poisoning in subadult mallards (Anas platyrhynchos) after oral administration included regurgitation, a reluctance to leave the swimming area, slow responses, muscular incoordination, heavy-footed walking, phonation, wing tremors, running and falling, weakness, and withdrawal (Hudson et al. 1984). Treatment concentrations as low as 3,300 µg/kg BW have produced signs of acrolein poisoning. These signs appeared as soon as 10 min after administration and persisted for as long as 36 days. At lethal oral concentrations, deaths occurred as soon as 32 min posttreatment and continued for several days (Hudson et al. 1984). Acrolein was lethal to developing avian embryos when whole eggs were injected with 51 to 182 µg/kg FW; in descending order, embryos were most sensitive when acrolein was administered by way of the yolk sac (51 μ g/kg), by the inner shell (82 µg/kg), and by the air sac (182 µg/kg; Table 3). Acrolein is 50 times more toxic to embryos of the domestic chicken (Gallus sp.) than acrylonitrile and 100 times more toxic than acrylamide (Kankaanpaa et al. 1979). Acrolein inhibits mucous transport in the trachea of the domes-

| Organism, route of administration, dose, and other variables | Effect | Reference ^a |
|---|---|------------------------|
| Mallard, Anas platyrhynchos; oral route; 9,100 μg/kg body weight (BW), 95% confidence interval [CI] of 6,300–13,100 μg/kg BW | LD50, age 3–5 months | 1 |
| Domestic chicken, Gallus sp. | | |
| Inhalation route | | |
| Adults subjected to 50,000 or 200,000 μg acrolein/L (113 or 454 mg/m³) air via an endotracheal cannula for up to 27 days | Decreases in trachea complement of ciliated and goblet cells; inhibited mucoous transport activity in trachea; lymphocytic inflammatory lesions in the tracheal mucosa. Changes were more pronounced at the higher dose and with increasing exposures | 2 |
| Air sac injection route. Embryos 2–3 days old; examined at day 13 | | |
| >127 µg/kg fresh weight (FW) whole egg | Dose-dependent decrease in survival | 3 |
| 182 μg/kg FW whole egg 1,818 μg/kg FW whole egg | LD50 LD80 | 3 3 |
| Air sac injection route. Embryos, 3-days old | | |
| 1 μg/kg FW whole egg | 20% developmental abnormalities vs. 5% in controls | 4 |
| 10 μg/kg FW whole egg 1,000 μg/kg FW whole egg | No malformations Lethal | 4 4 |
| Inner shell injection of membrane on heart route. Embryos, 72–76 h old; examined on day 14 of incubation | | |
| 25 μg/kg FW whole egg | No deaths or malformations | 5 |
| 51 μg/kg FW whole egg | 50% dead or malformed | 5 |
| 82 μg/kg FW whole egg 102 μg/kg FW whole egg | LD50 71% dead, 6% malformed | 5 5 |
| 203 μg/kg FW whole egg | 97% dead, 3% malformed | 5 |
| Yolk-sac injection route. Embryos 3-days old, examined at day 14 | | |
| 51 μg/kg FW whole egg | LD50 | 6 |
| 1,018 μg/kg FW whole egg | LD90; no evidence of increased teratogenicity over controls | 6 |

^a 1, Hudson et al. 1984; 2, Denine et al. 1971; 3, Chhibber and Gilani 1986; 4, Beauchamp et al. 1985; 5, Korhonen et al. 1983; 6, Kankaanpaa et al. 1979.

tic chicken (Denine et al. 1971), probably through ciliostatic action (EPA 1980). Adverse effects of acrolein were observed on chicken respiratory-tract physiology and pathology at greater than 50,000 $\mu g/L$ air (Table 3).

Malformations of the eye, coelom, neck, back, wings, and legs were observed in surviving

acrolein-treated chicken embryos (Korhonen et al. 1983) after whole eggs were injected with greater than 51 μ g acrolein/kg FW (Table 3). In other studies, acrolein showed no clear evidence of teratogenicity in chicken embryos, although there is a dose-dependent embryotoxic effect (Beauchamp et al. 1985; Chhibber and Gilani

1986). Acrolein-treated chicken embryos had a higher frequency of abnormal limbs, abnormal neck, and everted viscera than the controls, but the frequency was not dose-related. The overall incidence of abnormal embryos when treated at age 48 h was 24% but only 4% in controls; in embryos given acrolein at age 72 h, these values were 26% and 12% in controls (Chhibber and Gilani 1986).

Mammals

Acrolein is a strong cytotoxic and ciliostatic agent; its irritating effects on mucous membranes and its acute inhalation toxicity in mammals are well documented (Feron and Kruysse 1977; Feron et al. 1978; EPA 1980; Astry and Jakab 1983; Beauchamp et al. 1985; Leach et al. 1987; Leikauf et al. 1989). A characteristic of acrolein is its pungent, offensive, and acrid smell that is highly irritating to ocular and upper respiratory-tract mucosae (Beauchamp et al. 1985). Acrolein is toxic by all routes of exposure, and many of its toxic and biochemical effects are produced by interfering with critical sulfhydryl groups (Srivastava et al. 1992). In isolated rat-liver fractions, acrolein is a potent inhibitor of the high-affinity aldehyde dehydrogenase isozymes in mitochondrial and cytosolic fractions (Mitchell and Petersen 1988). Acrolein impairs DNA replication in vitro and inhibits certain mitochondiral functions (Feron et al. 1978). Studies with isolated rat livermembrane proteins revealed that acrolein inhibits plasma membrane enzymes and alters the membrane protein profile; this may be due to acrolein-induced polymerization of plasma-membrane proteins (Srivastava et al. 1992).

Measurable adverse effects of acrolein have been documented in representative species of mammals, but the severity of the effects are contingent on the mode of administration, concentration, dose, and duration of exposure (Table 4). Single oral doses of 4,000 µg/kg BW were lethal to guinea pigs and 28,000 µg/kg BW to mice; diets containing the equivalent of 500 µg/kg BW and more decreased survival in rats after 102 weeks (Table 4). Concentrations of 60,000 ug acrolein/L in drinking water had no measurable adverse effects on cows (Bos sp.) after 24 h; rats initially rejected drinking water containing 200,000 µg/L but eventually tolerated this concentration (Table 4). Dermal toxicity seems low; rabbits that were immersed up to their necks in water containing 20,000 µg acrolein/L for 60 min showed no adverse

effects (Table 4). No dermal sensitization occurred in healthy female guinea pigs (Cavia spp.) after repeated skin exposures to acrolein (Susten and Breitenstein 1990). In undiluted liquid or pungent vapor form, however, acrolein produces intense irritation of the eye and mucous membranes of the respiratory tract, and direct contact with the liquid can produce skin or eye necrosis (Beauchamp et al. 1985). A single intravenous injection of 850 ug acrolein/kg BW produced liver necrosis in rats; 6,000 µg/kg BW caused increased embryo resorption in mice (Table 4). Rats receiving near-lethal doses of acrolein by subcutaneous injection had liver and kidney damage and lung pathology (EPA 1980). Although subcutaneous injections revealed LD50 values between 164,000 and 1,022,000 µg/kg BW in rabbits, these results are questionable because acrolein may be sequestered at the injection site and delay delivery to the systemic circulation (Beauchamp et al. 1985). A single intraperitoneal injection of 1,000 µg/kg BW caused peritonitis in rats, and 7,000 µg/kg BW was lethal to mice; daily injections of 1.000 ug/kg BW were eventually lethal to rats (Table 4). Sublethal intraperitoneal injections of acrolein induced ascites, increased hematocrit, and prolonged sleeping times (Beauchamp et al. 1985). Acquired tolerance to acrolein in mice given repeated intraperitoneal injections suggests that an increased metabolism can partially explain the acquired tolerance (Warholm et al. 1984).

The largest number of studies of the toxicity of acrolein in animals was conducted by way of inhalation, probably because acrolein has an appreciable vapor pressure under ambient conditions and inhalation is the principal exposure for humans (Beauchamp et al. 1985). Because of their intolerance to sharp and offensive odor and to intense irritation of conjunctiva and the upper respiratory tract, humans have not suffered serious intoxication from acrolein. The strong lacrimatory effect of acrolein usually is a warning to occupational workers. Physiological perception of acrolein by humans begins at about 500 to 1,000 µg/L air with eye and nasal irritation; the irritating effects compel afflicted individuals to immediately leave the polluted area (Beauchamp et al. 1985). Laboratory animals died from inhalation of 8,000-11,000 μg/L after 4-6 h, mice from 875,000 μg/L after 1 min and rats from 660 µg/L after 24 days (Table 4). Animals dying from acute and subacute exposure to acrolein vapor had lung injury with hemorrhagic areas and edema (Albin 1962). Repeated

Table 4. Acrolein effects on selected mammals.

| administration, dose, and other variables | Effect | References |
|--|--|------------|
| Cow, Bos sp.; drinking water route; lactating dairy cows given 60,000 μg acrolein/L for 24 h | No change in feed or water intake or milk production; acrolein residues in milk <500 µg/L | 1 |
| Dog, Canis familiaris; inhalation route | | |
| 220, 1,000 or 1,800 µg/L air (0.5, 2.3, or 4.1 mg/m 3); continuous exposure for 90 days | Low concentration group appeared normal and gained weight. At 1,000 µg/L, ocular and nasal discharges. At the high concentration, severe irritation evident plus nonspecific inflammation of brain, heart, lung, liver, and kidney; no deaths | 2 |
| 400–600 μg/L air for 1–3 min | 81–84% of acrolein retained; accumulations greater in upper respiratory tract than lower respiratory tract | 3,4 |
| 700 or 3,700 μg/L air (1.6 or 8.4 mg/m³); exposure for 8 h daily, 5 days weekly for 6 weeks | Low concentration group appeared normal and gained weight. High concentration group visibly affected with weight loss, excessive salivation, ocular discharges, labored breathing, and histopathology of lung, liver, and kidney; blood and serum chemistry normal | 2 |
| 150,000 μg/L air (340 mg/m ³) for 30 min | LC50 | 3, 5, 6 |
| Guinea pig, Cavia spp.; inhalation route | | |
| 200, 1,000 or 1,800 μg/L continuously for 90 days | The low concentration group appeared normal. At 1,000 µg/L, pulmonary inflammation and liver necrosis. At high concentration, all had nonspecific inflammation of brain, heart, lung, liver, and kidney | 2 |
| 400 – $1,000 \mu g/L$ for $2 h$ | Decreased respiratory rate; effects reversed after exposure stopped | 6,7 |
| 400–1,000 μg/L for as long as 12 h | Concentration-related increases in respiratory resistance together with prolonged and deepened respiratory cycles | 3 |
| 700 or 3,700 μg/L; 8 h daily, 5 days weekly for 6 weeks | Low concentration group seemed normal. At high concentration, histopathology of lung, liver, and kidney | 2 |
| 10,500 μg/L for 6 h 20,000 μg/L for 10 min | LC50 Bronchioconstriction | 6 6 |
| Cat, Felis domesticus; inhalation route | | |
| 650,000 μg/L air for 2.25 h 870,000 μg/L air for 2.5 h | Died within 18 h Died during exposure | 6 6 |
| Human, Homo sapiens; inhalation route | | |
| 20 μg/L air | Threshold for affecting electrocortical activity | 6 |
| 30–40 μg/L air | Odor threshold for the most acrolein- sensitive people | 6 |

Table 4. Continued.

| Organism, route of administration, dose, | | |
|---|--|-------------------|
| and other variables | Effect | References |
| 90–300 μg/L air | Increasing concentration and increasing exposure caused increasing eye blinking, irritation, and decreasing respiratory frequency | 8 |
| 140–150 μg/L air for 2 min | Eye irritation in 30% of subjects | 6 |
| 250 μg/L air for 5 min | Moderate irritation of sensory organs | 3, 5 |
| 300 μg/L air for 10 min | Considerable acute irritation | 8 |
| 300–500 μg/L air | Odor threshold for most people | 3, 6 |
| 1,000 µg/L air for 1 min | Slight nasal irritation | 3, 5 |
| 1,000 μg/L for 5 min | Moderate nasal irritation; intolerable eye irritation | 3,5 |
| 1,800 µg/L air for 1 min | Slight eye irritation | 3 |
| 5,500 μg/L air for 20 sec | Painful eye and nasal irritation | 3, 5 |
| $21,800 \mu g/L$ air for $1 sec$ | Intolerable | 3, 5 |
| Syrian golden hamster, Mesocricetus auratus | | |
| Gavage route; 1,000 μg/animal, equivalent to about 4,000 μg/kg BW | Fatal within a few hours | 11, 12 |
| Inhalation route | • | |
| 400 or 1,400 μg/L air; exposure for 6 h daily, 5 days weekly for 13 weeks | No adverse effects at low concentration; nasal histopathology at high concentration | 9 |
| 4,000 μg/L air (9.2 mg/m³); 7 h daily, 5 days weekly for 52 weeks | No effect on survival; no indication of cancer. Abnormal behavior, growth retardation, increased lung weight, decreased liver weight, nasal histopathology | 10 |
| $6,000 \mu g/L air (13.8 mg/m^3) for 4 h$ | Cytotoxic to airway cells | 3 |
| 25,400 μg/L air for 4 h | LC50 | 6, 10 |
| Iouse , <i>Mus</i> sp. | | |
| Drinking water route; $24,000 \mu g/L$ for 18 months | Death | 29 |
| Inhalation route | | |
| 10 μg/L air continuously for | Some reduction in | 4 |
| 5 weeks | pulmonary compliance | |
| 1,000–2,000 µg/L (2.3–4.6 mg/m³) air for 24 h | Decreased pulmonary ability to kill bacteria Staphylococcus aureus and Proteus mirabilis | 3 |
| 1,700 μg/L air for 10 min | 50% reduction in respiratory rate | 6,14 |
| 3,000 or 6,000 μg/L air for 8 h | Concentration-dependent impairment of pulmonary antibacterial responses | 15 |
| 6,000–15,000 μg/L air; 6 h daily, | Decreased body weight (6%) in all test | 3 |
| 5 days weekly for 6 weeks | groups, but not concentration-related | ~ |
| 66,000 μg/L for 6 h 175,000 μg/L air for 10 min | LC50, 24 h post-exposure LC50 | 256 |
| 875,000 μg/L air for 10 mm | LC50 | 3, 5, 6 $3, 5, 6$ |
| Intraperitoneal injection route | | • |
| 4,000 μg/kg BW; single injection | Plasma total lactic dehydrogenase activity (LDH) increased 5 times, | 16 |
| 4,000 μg/kg BW; multiple daily or weekly injections | with peak after 10 h Progressively less pronounced effect on LDH activity | 16 |

Table 4. Continued.

| Fable 4. Continued. | | |
|--|--|--------------------------------------|
| Organism, route of administration, dose, and other variables | Effect Re | ferences ^a |
| 7,000 µg/kg BW; single injection 12,000 µg/kg BW; preceded by daily injections of 4,000 µg/kg BW for 5 days | LD50 50% mortality | 16 16 |
| Oral route; 28,000 µg/kg BW | Acute oral LD50 | , 4, 5, 6 |
| Rabbit, Oryctolagus sp. | | |
| Dermal route; immersed up to necks for 60 min in water with 20,000 μ g/L | No adverse effects | 1 |
| Drinking water route | | |
| 9,000 µg/L for 13 days 36,000 µg/L for 13 days | Miscarriages Stomach ulcers | 29 29 |
| Inhalation route | | |
| 400 or 1,400 μg/L; 6 h daily, 5 days weekly for 13 weeks | No adverse effects at low concentration; some signs of distress at 1,400 µg/L | 9 |
| 600 μg/L; 4 h daily for 30 days 1,700–2,400 μg/L for 10 min; with or without 1,000 μg ozone/L | No ocular effects Acrolein alone had no effect on respiratory rate. Ozone-acrolein mixtures produced a marked decrease in respiratory rate | 6 |
| 4,900 μg/L air; 6 h daily, 5 days weekly for 13 weeks | Ocular and nasal irritation, growth depression, respiratory tract histopathology | 10 |
| 6,500–10,500 μg/L; exposure duration unknown | Emphysema, tracheobronchitis, some deaths | 2 |
| 10,500 μg/L air for 6 h | LC50 | 6 |
| Intravenous injection route; 3,000, 4,500 or 6,000 ug acrolein/kg BW on day 9 of gestation; killed on day 28 of gestation Percutaneous injection route | Embryo resorption was significantly higher in 6,000 μg/kg group vs. controls, but was the same as controls in lower concentration groups | 6 |
| 164,000 μg/kg BW 238,000 μg/kg BW 335,000 μg/kg BW 562,000 μg/kg BW 1,022,000 μg/kg BW | LD50 for 20% acrolein in mineral spirits LD50 for 10% acrolein in mineral spirits LD50 for 20% acrolein in water LD50 for undiluted acrolein LD50 for 10% acrolein in water | 3, 5 3, 5 3, 5 3, 5 3, 5 |
| Domestic sheep, Ovis aries; inhalation route via cervical trachea; ewes, 3-4 years old; exposed to smoke containing high (but unknown) concentrations of acrolein for 20 min; killed 1-22 days after exposure | Within 24 h of exposure there was sloughing of total cervical tracheal epithelium and a 35% reduction in tracheal basal cells; trachea was normal 18–22 days after exposure | 13 |
| Baboon, Papio anubis; inhalation route. Juveniles exposed to air concentrations of 12,000–2,780,000 µg acrolein/L (272–63,100 mg/m³) for 5 min, then tested for learned avoidance/escape response | Avoidance/escape response enhanced in all animals at all concentrations tested. The group exposed to 1,025,000 µg/L air died wit respiratory complications within 24 h post-exposure. The group exposed to the highest concentration of 2,780,000 µg/L for 5 min died within 90 min postexposure with severe respiratory complications | |

Table 4. Continued.

| Organism, route of administration, dose, and other variables | Effect | References ^a |
|---|---|-------------------------|
| Laboratory white rat, Rattus sp. | 221000 | Telefences |
| Dermal route; exposure duration and dose unknown | Skin burns; severe ocular effects | 18 |
| Drinking water route | | |
| 5,000, 13,000, 32,000, 80,000 or 200,000 μg/L for 12 weeks | Water consumption in the 200,000 µg/L group was reduced by about 33% for the first 3 weeks; by week 12, all groups appeared normal and had apparently adapted to the odor and taste of acrolein | |
| 80,000 μg/L for 3 days | Some deaths | 29 |
| 100,000 or 250,000 μg/L for 124 weeks | No increase in tumors over controls; | 11 |
| 124 weeks 100,000, 250,000 or 625,000 μg/L for 120 weeks | no decrease in survival No significant decrease in survival when compared to controls. The 100,000 μg/L group had a 30% frequency of liver neoplasms and a 5% frequency of adrenal cortex neoplasms; however, no neoplasms were found in the 250,000 μg/L group. The 625,000 μg/L group had a 10% frequency of liver neoplasms vs. 25% in controls | 12 |
| 200,000 μg/L for 90 days | No adverse effects | 1 |
| 600,000, 1,200,000 or 1,800,000 μg/L for 60 days | Rats in the two high-concentration groups refused to drink and all died, apparently from dehydration. In the low-concentration group 20% died, but survivors were not dehydrated and had no tissue pathology | 4 |
| 625,000 μg/L for 100 weeks | No decrease in survival; 20% of females developed adrenal cortical ademonas and 10% had neoplastic nodules in the adrenal cortex vs. 0% in controls | 6 |
| 625,000 μ g/L for 104 weeks | No decrease in survival. Increased frequency of adrenal cortex adenomas in females: 25% vs. 1.3% in controls | 11 |
| Inhalation route | | |
| 10 or $50~\mu\text{g/L}$ air for $1~\text{min}$ | Increased blood pressure and heart beat rate | 19 |
| 10, 500, 1,000, or 2,400 μg/L air for 3 h | At 500 µg/L and higher, effects on respiratory mucosa included depletion of nonprotein sulfhydryl (NPSH) concentration and slight decrease in protein sulfhydryl (PSH) concentration. Effects on olfactory mucosa showed no changes in PSH at all test concentrations, but significant depletion of NPSH in the two high-concentration groups | 20 |

Table 4. Continued.

| Organism, route of administration, dose, and other variables | Effect | References ^a |
|---|--|-------------------------|
| 100, 1,000, or 3,000 μg/L air, exposed 6 h daily, 5 days weekly for 3 weeks | No adverse effects in the two low- concentration groups. The high concentration group had | 14 |
| | depressed spleen weight and body weight and extensive nasal histopathology | |
| 150, 510, or 1,520 μg/L air; continuous exposure for 61 days | At low concentration, no respiratory tract lesions or deaths. At 510 µg/L, bronchial epithelium abnormalities but all survived. At high concentration, reduced survival; bronchopneumonia and bronchial abnormalities in survivors | 6 |
| 220 or 660 μg/L air; exposed continuously for 60 days | No deaths at 220 μ g/L; 70% died within 24 days at 660 μ g/L | 2 |
| 220, 1,000 or 1,800 μg/L air, exposed continuously for 90 days | The low concentration group appeared normal and gained weight. At 1,000 µg/L, liver necrosis and pulmonary hemorrhage. At 1,800 µg/L, all had nonspecific inflammation of brain, heart, lung, liver, and kidney | 2 |
| 400 μg/L air; exposed 6 h daily, 5 days weekly for 13 weeks | Nasal histopathology | 9 |
| 400, 1,400, or 4,000 μg/L air; exposed 6 h daily, 5 days weekly for 62 days | Some bronchial histopathology at 1,400 µg/L; some deaths among males at 4,000 µg/L | 6 |
| 520 μg/L (1.2 mg/m ³); continuous exposure for 30 days | Decreased growth, altered liver enzyme activity | 3 |
| 550 ug/L air; continuous exposure for up to 77 days | Upper respiratory irritation, reduced resistance to infection by Salmonella, and increased pulmonary macrophages; all effects disappeared by day 63 | 6 |
| 700 or 3,700 μg/L air; exposed 8 h daily, | No adverse effects noted at low concentration. At 3,700 µg/L, | 2 |
| 5 days weekly for 6 weeks | histopathology of lung, liver, and kidney | |
| $2,000~\mu g/L$ air for $40~h$ | Increased hepatic alkaline phosphatase activity; increased liver and adrenal weight | 3 |
| $2,500$ – $5,000~\mu g/L$ air for 1 min | Cardioinhibitory effect that was reversed within 10 sec after inhalation of acrolein ceased | 19 |
| 4,900 μg/L air; exposed 6 h daily, 5 days weekly for 13 weeks | 50% mortality during first 4 weeks with no deaths thereafter. Survivors had depressed growth and respiratory tract histopathology | 9 |
| 6,000–8,888 μg/L air; exposed 6 h daily, 5 days weekly for 3 weeks | Most died within 5 exposure days | 14 |
| 8,000–8,300 μg/L air for 4 h | LC50 within 14 days; death due to lung injury | 5, 6, 21 |

Table 4. Continued.

| administration, dose, | TACC 1 | D.c 5 |
|---|--|------------|
| and other variables | Effect | References |
| 26,000 μg/L air for 1 h | LC50 | 21 |
| 43,500–304,000 μg/L air for 30 min | Respiratory distress; nasal and ocular irritation; some deaths in 4–5 days; pulmonary | 2, 6 |
| | edema; bronchial degeneration; excess blood in heart, liver, and kidney | |
| 131,000 μg/L air for 30 min | 50% dead in 3 weeks | 6 |
| 283,000 μg/L air; daily 10-min exposures for 6 months | No deaths; some bronchial pathology | 6 |
| 326,000 µg/L air; daily 10-min exposures for 6 months | 50% dead; tracheobronchial pathology | 6 |
| 435,000 μg/L air; daily 10-min exposures for 6 months | All dead; severe histopathology of respiratory tract | 6 |
| 5,000,000–10,000,000 μg/L air for 5 min | Rats on a motor-driven exercise wheel were incapacitated within 5–7 min and died shortly thereafter | 17 |
| Intraamniotic injection route Embryos given 0.1, 1, 10, or 100 µg of acrolein on day 13 of gestation; | 98–100% dead at 10 and 100 µg; 85% of live fetuses receiving 1 µg were malformed (edema, hydrocephaly, cleft palate, defects | 6 |
| examined on day 20 of gestation | of limbs and tail); no teratogenic effects at $0.1~\mu g$ | |
| Intraperitoneal injection route | | |
| 1,000 μg/kg BW, single injection | Peritonitis | 22 |
| 1,000 µg/kg BW daily for at least 5 days | Lethal | 22 |
| 2,000 μg/kg BW twice a week for 6 weeks followed by uracil as 3% of the diet for 20 weeks, then control diet for | Acrolein followed by uracil produced a 60% incidence of papilloma in urinary bladder in treated group (acrolein plus uracil) vs. 27% in | 22 |
| 6 weeks | water controls (uracil only). No | |
| 0.500/I DVV 1. 11 | tumors in either group | |
| 2,500 μg/kg BW daily 3,360 μg/kg BW; single injection; | All dead after second dose Most (89%) of the acrolein recovered | 3 |
| tissues analyzed after 24 h | was in the acid-soluble fraction of the liver, 3% in the liver lipids, and minor amounts (0.4–1.7%) in liver proteins and RNA and DNA fractions | 3 |
| Intravenous injection route | | |
| 50-500 μg/kg BW to | At 50-200 µg/kg BW, blood pressure | 23 |
| spontaneously hypersensitive rats | increased; at 300–500 µg/kg BW, blood pressure decreased | 20 |
| 250–1,000 μg/kg BW | Increased blood pressure within 5 sec which peaked at 20–30 sec and lasted about 1 min | 19 |
| 850 or 1,700 μg/kg BW | Liver necrosis | 3 |
| 10,000 μg/kg BW | Cardioinhibitory effects | 19 |

Table 4. Continued.

| and other variables | | TO 0 0 |
|--|--|----------------------|
| and other variables | Effect | References |
| In vitro studies | | |
| Cultured embryos | | |
| 4,500 μg/L serum medium | Growth retardation; 50% malformation frequency among survivors | 24, 25 |
| 6,700 μg/L serum medium | Mortality of 64%; all surviving embryos malformed | 24, 25 |
| 7,800–9,000 μg/L serum medium | All dead | 24, 25 |
| 160 μg/L serum-free medium | 50% frequency of malformations in brain, facial area, and heart | 25 |
| 300–1,100 μ g/L serum-free medium | 50%–100% lethal | 25 |
| Cultured myocytes and fibroblasts from neonatal heart | | |
| 600 µg/L culture medium for 4 h | Myocyte ATP levels reduced | 26 |
| 2,800 μ g/L culture medium for 4 h | Irreversible cell lysis and ciliostasis | |
| Isolated liver fractions | | |
| 1,700 μg/L medium | Mitochondrial aldehyde dehydrogenase (ALDH) activity inhibited 91%; cytosolic ALDH activity inhibited 33% | 27 |
| 2,700 μg/L medium; 5 sec preincubation in aldehyde substrate | Inhibition of mitochondrial and cytosolic ALDH | 27 |
| Oral route | | |
| Daily gavage of 50, 500, or 2,500 μg/kg BW for 102 weeks | Dose-related mortality in males during first year and in females during entire study; significant lethality in the 500 and 2,500 µg/kg groups. No increased incidence of microscopic neoplastic or nonneoplastic lesions in treated rats; decreased creatinine | |
| Two treatments of 4,000–10,000 µg/kg BW (estimated), 2–3 days apart; total dose of 8,000–20,000 µg/kg BW | phosphokinase levels in treated rats All died shortly after the second dose | 12 |
| 5,000 µg/kg BW daily for 9 days via stomach intubation | No deaths | 3 |
| 10,000 µg/kg BW, single stomach intubation | Fatal | 3 |
| 25,000 µg/kg BW, single gastric dose | LD50 within 48 h | 22 |
| 42,000–46,000 μg/kg BW, single dose | LD50 within 14 days | 3, 4, 5, 6, 18,22 |
| Squirrel monkey, Saimiri sciurea; inhalation route | | |
| Illianation route | | |

Table 4. Continued.

| Organism, route of administration, dose, and other variables | Effect | References |
|--|---|------------|
| continuous exposure for 90 days | normal and gained weight; 1,000 µg/L monkeys were visibly affected with ocular and nasal discharges. No deaths at 1,800 µg/L, but excessive salivation, ocular discharges, and hyperplasia of trachea | |
| 700 or 3,700 μg/L air; 8 h daily, 5 days weekly for 6 weeks | Low concentration group appeared normal. High dose group had weight loss; histopathology of lung, liver, and kidney; 22% mortality (2 of 9 died on days 6 and 9 of exposure) excessive salivation, and frequent blinking | |

^a 1, Ferguson et al. 1961; 2, Lyon et al. 1970; 3, EPA 1980; 4, NRC 1977; 5, Albin 1962; 6, Beauchamp et al. 1985; 7, Leikauf et al. 1989; 8, Weber-Tschopp et al. 1977; 9, Feron et al. 1978; 10, Feron and Kruysse 1977; 11, Lijinsky and Reuber 1987; 12, Lijinsky 1988; 13, Barrow et al. 1992; 14, Leach et al. 1987; 15 Astry and Jakab 1983; 16, Warholm et al. 1984; 17, Kaplan 1987; 18, Sine 1991; 19, Egle and Hudgins 1974; 20, Heck et al. 1986; 21, Ballantyne et al. 1989; 22, Cohen et al. 1992; 23, Green and Egle 1983; 24, Slott and Hales 1987; 25, Slott and Hales 1986; 26, Toraason et al. 1989; 27, Mitchell and Petersen 1988; 28, Parent et al. 1992; 29, ATSDR 1990.

exposures of hamsters, rats, and rabbits to high sublethal concentrations of acrolein caused ocular and nasal irritation, growth depression, and respiratory tract histopathology in all species (Feron and Kruysse 1977; Table 4). However, repeated exposures to low, tolerated concentrations of acrolein did not produce toxicological effects (Albin 1962), suggesting that acrolein effects are not cumulative and that minimal damage is quickly repaired.

Inhaled acrolein—in ug acrolein/L air—had sublethal effects at 10-50 for 1 min on rats (increased blood pressure and heart rate); at 10 for 5 weeks on mice (reduction in pulmonary compliance); at 140-150 for 2 min on humans (eye irritation in 30%); at 300-500 on humans (odor threshold); at 300 for 10 min on humans (acute irritation); at 400 for 13 weeks on rats (nasal histopathology); at 400-600 for 1-3 min on dogs (accumulations in upper respiratory tract); and at 1,000 for 90 days on dogs, monkeys, and guinea pigs (ocular and nasal discharges; Table 4). Sublethal effects of inhaled acrolein in representative small laboratory mammals were greatest on the upper respiratory tract and bronchial airways and included edema, ciliastasis, inflammation, degenerative loss of epithelia, altered ventilatory function, and bronchoconstriction (Feron and Kruysse 1977; Feron et al. 1978; EPA 1980; Astry and Jakab 1983; Beauchamp et al. 1985; Barkin et al. 1986; Leach et al. 1987; Leikauf et al. 1989; Table 4). Typical signs of toxicity from inhaled acrolein in small mammals include ocular and nasal irritation; growth depression; shortness of breath; lesions in the urinary tract, respiratory tract, trachea, and nasal passages: laryngeal edema; reduced resistance to bacterial infection; enlarged liver and heart; elevated blood pressure and heart rate; altered enzyme activities; and protein synthesis inhibition (EPA 1980; Beauchamp et al. 1985; Leach et al. 1987; Table 4). Signs of inhaled acrolein toxicity varied significantly with dose and species. For example, acrolein toxicity in rats at environmental concentrations was confined to local pathologic nasal changes, including metaplastic, hyperplastic, and dysplastic changes in the mucous, respiratory, and olfactory epithelium of the nasal cavity (Leach et al. 1987). Some inhaled toxicants, including acrolein, can prolong bacterial viability in the lung and thus enhance severeness of the disease. Mice convalescing from viral pneumonia became severely deficient in antibacterial defenses when exposed to acrolein (Astry and Jakab 1983). But acrolein-treated mice subjected to 100 µg/L air (5 consecutive daily 3-h exposures) were not significantly sensitive to pulmonary bacteria Klebsiella pneumoniae or Streptococcus zooepidemicus (Aranyi et al. 1986).

Acrolein may be a carcinogen, cocarcinogen, or tumor initiator. As an aldehyde with strong affinity to sulfhydryl groups, acrolein is theoretically

expected to remove free tissue thiols—compounds that protect bronchial epithelia against attack by carcinogens (Feron and Kruysse 1977; Feron et al. 1978). Carcinogenicity from inhalation of acrolein has not been reported (Lijinsky and Reuber 1987), and acrolein was not an evident cofactor in studies of respiratory-tract carcinogenesis with hamsters (Cricetus spp.) exposed to benzo(a)pyrene or diethylnitrosamine (Feron and Kruvsse 1977). Moreover, long-term studies with rodents given acrolein by gavage did not increase incidences of neoplastic or nonneoplastic lesions (Parent et al. 1992). Other studies, however, suggest that acrolein is carcinogenic. Compounds closely related to acrolein are carcinogenic to rodents and humans and include acrylonitrile (vinyl cyanide) and vinyl acetate (Lijinsky 1988). Glycidaldehyde—an acrolein intermediate metabolite—is classified as an animal carcinogen by The International Agency for Research on Cancer; however, no convincing data are available on the carcinogenic potential of acrylic acid and other acrolein metabolites (Beauchamp et al. 1985). Acrolein at least partially can account for the initiating activity of cyclophosphamide carcinogenesis (Cohen et al. 1992). Cyclophosphamide and its analogs are a group of chemotherapeutic and immunosuppressive drugs; toxic side effects of this drug group are attributed to its metabolites, especially acrolein (Cohen et al. 1992). Acrolein is a suspected carcinogen because of its 2,3-epoxy metabolite and its weak mutagenic activity in the Salmonella screen (Leach et al. 1987). Acrolein may be a weak carcinogen, as judged by the increased frequency of adrenal adenomas in female rats after exposure for 2 years to drinking water with 625,000 µg acrolein/L (Lijinsky and Reuber 1987). Acrolein has cancer-initiating activity in the rat urinary bladder, but studies with N-[4-(5-nitro-2furyl)-2 thiazoyl] formamide precluded evaluation of acrolein as promoting a complete carcinogenic activity from low rodent survival (Cohen et al. 1992). Additional studies seem needed to evaluate the carcinogenic potential of acrolein.

After intraamniotic injection, acrolein is teratogenic to rats in vivo but not in vitro. When administered intraamniotically to the whole embryo culture system of the rat on day 13 of gestation, acrolein caused edema, hydrocephaly, open eyes, cleft palate, abnormal umbilical cord, and defects of the limbs and face (Slott and Hales 1986). Beauchamp et al. (1985) suggest that acrolein-associated teratogenicity is caused by acrylic acid,

an acrolein metabolite. Acrylic acid injected into amniotic fluid of rats on day 13 of gestation produced a dose-dependent increase in the percentage of fetuses with skeletal and other abnormalities (Beauchamp et al. 1985).

Acrolein can react synergistically, additively, or antagonistically with other chemicals (Beauchamp et al. 1985). Rat embryos were protected by glutathione against acrolein-induced mortality. growth retardation, and developmental abnormalities-provided that glutathione was concurrently present with acrolein. When rat embryos were cultured in the presence of acrolein for 2 h prior to glutathione exposure, there was no protection against acrolein-induced embryolethality, teratogenicity, and growth retardation (Slott and Hales 1987). Acrolein effects—including altered liver-enzyme activity in rats—were reduced by pretreatment of animals with chemicals that inhibited protein synthesis (NRC 1977). Exposure to acrolein is sometimes accompanied by exposure to formaldehyde and other short-chain saturated aliphatic aldehydes, which in combination cause allergic contact dermatitis (Susten and Breitenstein 1990). A 40-mL puff of cigarette smoke contains 8.2 µg of acrolein and 4.1 µg of formaldehyde; irritation, ciliastasis, and pathologic changes of the respiratory tract from both compounds have been widely studied (Egle and Hudgins 1974). The toxicities of acrolein and formaldehyde seem similar; both exert their principal effects in the nasal passages (Leach et al. 1987). Acrolein in combination with formaldehyde was synergistic in reducing respiratory rates in mice; however, mixtures of sulfur dioxide and acrolein were antagonistic (Beauchamp et al. 1985). Formaldehyde pretreatment (15,000 μ g/L, 6 h daily for 9 days) of rats protects against respiratory-rate depression by acrolein. Rats pretreated with formaldehyde had a 50% respiratory-rate depression at 29,600 µg acrolein/L versus 6,000 µg/L from acrolein alone (Babiuk et al. 1985), suggesting cross tolerance. Effects of interaction of acrolein with other toxicants are not comparable between rodents and humans. In rodents, the presence of irritant gases in smoke-such as acrolein-may delay the effects of other toxicants. In humans, however, the inhalation of acrolein and other irritant gases may cause a hypoxemic effect that can enhance the effects of hypoxia-producing gases (Kaplan 1987).

Some chemicals normally contain acrolein as a metabolite or impurity. For example, allylamine toxicity to the rat cardiovascular system is be-

lieved to involve metabolism of allylamine to the highly reactive acrolein (Toraason et al. 1989). Certain mercapturic acids can be used as biological markers of exposure for chemicals that are metabolized to acrolein and excreted as mercapturic acid in the urine (Sanduja et al. 1989). In one case, rats given 13,000 µg acrolein/kg BW by gavage excreted 79% of the acrolein and 3-hvdroxypropylmercapturic acid (3-OHPrMCA) in urine within 24 h. These data suggest that 3-OHPrMCA can be used as a marker of exposure to allylic and other compounds that lead to the formation of acrolein (Sanduja et al. 1989). The common industrial chemical MDP (2-methoxy-3.4-dihydro-2PH-pyran) is frequently contaminated with acrolein during its synthesis; MDP causes severe irritancy and death of rats from accumulation of acrolein vapor (Ballantyne et al. 1989). Sparging acrolein-contaminated MDP with nitrogen gas before atmospheric release significantly reduced or abolished lethal toxicity to rats (Ballantyne et al. 1989).

Recommendations

Agricultural crops can usually tolerate as much as 15,000 µg of acrolein/L of irrigation water: however, aquatic invertebrates and fish die in acute exposures to 55-68 µg/L or in chronic exposures to greater than 21 µg/L (Table 5). Those who use acrolein to control submerged aquatic macrophytes are strongly advised that acrolein treatment at recommended application concentrations also eliminates nontarget fish and aquatic invertebrates. No acrolein criteria are now available or promulgated by regulatory agencies for the protection of avian and terrestrial wildlife; this seems to be a high-priority research need. Beauchamp et al. (1985) recommend additional research in

several areas: long-term effects of acrolein inhalation on carcinogenicity and respiratory histology with rodent models; biochemical mechanisms of acrolein toxicity; genotoxic potential with chromosome breakage and exchange systems; acute and chronic toxicity from interaction effects of acrolein with other gases; and fate of accumulated acrolein in animals.

The human threshold concentration of acrolein in the United States for an 8-h workday and 40-h workweek is 110 µg/L (0.25 mg/m³) air; the shortterm exposure limit is 350 µg/L (0.8 mg/m³) air and is predicated on continuous exposure of workers for short intervals (Table 5; Beauchamp et al. 1985). Humans can tolerate a total daily intake of 47.8 µg of acrolein, equivalent to 0.68 µg/kg BW by a 70-kg individual (Table 5).

For handling acrolein, gloves, vapor-proof goggles or a full-face mask, and other protective clothing are mandatory (Albin 1962; Beauchamp et al. 1985; NIOSH 1990). Acrolein spills should be neutralized with 10% sodium bisulfite solutions (Albin 1962). Air packs or fresh-air breathing masks, safety showers, and eye baths should be available wherever acrolein is handled (Beauchamp et al. 1985). Purging confined areas with nitrogen is recommended prior to entering a suspected acrolein-contaminated enclosure. The eyes are particularly susceptible to liquid acrolein and, if exposed, should receive prompt treatment, although severe residual injury is probable regardless of treatment; dilute solutions of acrolein may also cause residual eye injury. Acrolein represents a serious fire hazard because of its high flammability and potential for vapors to form explosive mixtures with air. Flame-proof electrical equipment and proper grounding is required to prevent acrolein ignition. Individuals exposed to acrolein

Table 5. Proposed acrolein criteria for the protection of living resources and human health.

| Resource, criterion, and other variables | Concentration | Reference ^a |
|--|---|------------------------|
| Agricultural Crops | | |
| Irrigation water, tolerated level | <15,000 μg/L | 1 |
| Aquatic life | | |
| Freshwater organisms | | |
| Sensitive species, tolerated level | | |
| Acute exposures | <68 μg/L | 2 |
| Chronic exposures | <21 μg/L | 2 |
| Rainbow trout, safe level | 20 μg/L for <48 h or 200 μg/L for <4.8 h | 3 |

Table 5. Continued.

| Resource, criterion, and other variables | Concentration | Reference ^a |
|---|---|------------------------|
| Marine organisms; acute exposures, tolerated level | <55 μg/L | 2 |
| Laboratory white rat | | |
| Air | | |
| Maximum daily average Maximum daily | <13 μ g/L (<0.03 mg/m ³) <44 μ g/L (<0.1 mg/m ³) | 7 7 |
| Human health | | |
| Air | | |
| Maximum allowable emission concentration in populated areas of former Soviet Union | 132 μg/L (0.3 mg/m 3) | 4 |
| No observable effect level 90-day confined space (i.e., submarines) guideline | <22 μg/L (<0.05 mg/m ³) 22 μg/L (0.05 mg/m ³) | 4 5 |
| Odor threshold | $<44 \mu g/L (<0.1 \text{ mg/m}^3)$ | 4 |
| Maximum acceptable concentration in room air of former Soviet Union | 44 μ g/L (0.1 mg/m ³) | 2,4 |
| Irritation threshold mg/m ³) | 44–88 μ g/L (0.1–0.2 m g/m ³) | 4 |
| Occupational exposure standard (8 h daily, 40 h work week) in United States; not to exceed in most European countries, Australia, and Japan | 100–110 μg/L (0.25 mg/m ³) | 2, 4, 5, 6, 8 |
| Occupational exposure standard in Hungary and former Soviet Union | 308 μg/L (0.7 mg/m ³) | 4 |
| Maximum 15-min exposure limit in USA workplace | $300-352 \mu g/L (0.8 mg/m^3)$ | 4, 8 |
| Ceiling standard for occupational exposure in the former Czechoslovakia | 440 μg/L (1.0 mg/m ³) | 4 |
| Acceptable ambient air concentrations | | |
| New York | $0.83 \mu \text{g/m}^3$ for 1 year | 9 |
| Florida | 2.5 μg/m ³ for 8 hr | 9 |
| North Dakota | 8.0 μ g/m ³ for 1 hr 80 μ g/m ³ for 15 min | 9 9 |
| North Carolina | оо на пот 10 пот | - |
| Diet Water plus consumption of | <320 μg/L medium | 2 |
| contaminated aquatic organisms from that water body | | |
| Consumption of contaminated aquatic organisms alone | <780 μg/L medium | 2 |
| Food packaging materials; food starch | <0.6% | |
| Total daily intake | $<47.8 \mu g = <0.68 \mu g/kg$ body weight daily for a 70-kg person | . 2 |

^a 1, Ferguson et al. 1961; 2, EPA 1980; 3, Bartley and Hattrup 1975; 4, Beauchamp et al. 1985; 5, Lyon et al. 1970; 6, Leach et al. 1987; 7, NRC 1977; 8, NIOSH 1990; 9, ATSDR 1990.

by inhalation should be removed from the area and given oxygen; subsequent treatment by physicians of pulmonary inflammation with corticosteroids and hydroxocobalamin is recommended even if there are no symptoms (Beauchamp et al. 1985) because adverse effects from acrolein exposure may not become apparent until 4-24 h after exposure (Albin 1962). Oxygen therapy should be continued and analgesics given for relief of other symptoms as necessary (Beauchamp et al. 1985). There are many synthetic and natural sources of acrolein; however, special precautions are recommended when acrolein occurs as a contaminant in the synthesis of widely used chemicals such as 2-methyoxy-3,4-dihydro-2H pyran (Ballantyne et al. 1989).

Acknowledgments

I thank L. Garrett and W. Manning for library services; M. Holmes for secretarial assistance; and S. J. Hamilton, P. F. P. Henry, O. H. Pattee, and several anonymous reviewers for scientific and technical reviews of the manuscript.

Cited Literature

- Albin, T. B. 1962. Handling and toxicology. Pages 234– 239 in C. W. Smith, editor. Acrolein. John Wiley, New York.
- Anderson, E. A., and G. C. Hood 1962. Physical properties. Pages 7–50 in C. W. Smith, editor. Acrolein. John Wiley, New York.
- Agency for Toxic Substances and Disease Registry (ATSDR). 1990. Toxicological profile for acrolein. U.S. Public Health Service, TP-90-01. 145 pp.
- Aranyi, C., W. J. O'Shea, J. A. Graham, and F. J. Miller. 1986. The effects of inhalation of organic chemical air contaminants on murine lung host defenses. Fundamental and Applied Toxicology 6:713–720.

Astry, C. L., and G. J. Jakab. 1983. The effects of acrolein exposure on pulmonary antibacterial defenses. Toxicology and Applied Pharmacology 67:49–54.

- Babiuk, C., W. H. Steinhagen, and C. S. Barrow. 1985. Sensory irritation response to inhaled aldehydes after formaldehyde pretreatment. Toxicology and Applied Pharmacology 79:143-149.
- Ballantyne, B., D. E. Dodd, I. M. Pritts, D. J. Nachreiner, and E. H. Fowler. 1989. Acute vapour inhalation toxicity of acrolein and its influence as a trace contaminant in 2-methyoxy-3, 4-dihydro-2H-pyran. Human Toxicology 8:229-235.
- Barkin, P., W. Jung, P. Pappagianopoulos, C. Balkas,
 D. Lamborghini, J. Burke, and C. Hales. 1986. The
 role of the bronchial circulation in production of pul-

- monary edema in dogs exposed to acrolein in smoke. The American Review of Respiratory Disease 133:A270.
- Beauchamp, R. O., Jr., D. A. Andjelkovich, A. D. Kligerman, K. T. Morgan, and H. d'A. Heck. 1985. A critical review of the literature on acrolein toxicity. CRC Critical Reviews in Toxicology 14:309–380.
- Baron-Marano, F. L., and M. C. Izard. 1968. Observation d'anomalies ultrastructurales dans la descendance d'algues traitées par l'acroléine. Comptes Rendues Hebdomadaires des Séances de l'Académie des Sciences. D, Sciences Naturelles 267:2137-2139.
- Barrow, R. E., C-Z. Wang, R. A. Cox, and M. J. Evans. 1992. Cellular sequence of tracheal repair in sheep after smoke inhalation injury. Lung 170:331–338.
- Bartley, T. R., and A. R. Hattrup. 1975. Acrolein residues in irrigation water and effects on rainbow trout. U.S. Department of the Interior, Denver Engineering and Research Center, Bureau of Reclamation Report REC-ECR-75-8. 11 pp.
- Bowmer, K. H., and M. L. Higgins. 1976. Some aspects of the persistence and fate of acrolein herbicide in water. Archives of Environmental Contamination and Toxicology 5:87-96.
- Bowmer, K. H., and G. R. Sainty. 1977. Management of aquatic plants with acrolein. Journal of Aquatic Plant Management 15:40–46.
- Bowmer, K. H., G. R. Sainty, G. Smith, and K. Shaw. 1979. Management of *Elodea* in Australian irrigation systems. Journal of Aquatic Plant Management 17:4–12.
- Bowmer, K. H., and G. H. Smith. 1984. Herbicides for injection into flowing water: acrolein and endothalamine. Weed Research 24:201–211.
- Burdick, G. E., H. J. Dean, and E. J. Harris. 1964. Toxicity of aqualin to fingerling brown trout and bluegills. New York Fish and Game Journal 11:106–114.
- Chhibber, G., and S. H. Gilani. 1986. Acrolein and embryogenesis: an experimental study. Environmental Research 39:44–49.
- Cohen, S. M., E. M. Garland, M. St. John, T. Okamura, and R. A. Smith. 1992. Acrolein initiates rat urinary bladder carcinogenesis. Cancer Research 52:3577-3581.
- Comendador, M. A. 1984. Variation of sensitivity to acrolein during the development of *Drosophilia melanogaster*. Brazilian Journal of Genetics 7:411-417.
- Corbus, F. G. 1982. Aquatic weed control with endothall in a Salt River project canal. Journal of Aquatic Plant Management 20:1–3.
- Denine, E. P., S. L. Ribbins, and C. J. Kensler. 1971. The effects of acrolein inhalation on the tracheal mucosa of the chicken. Toxicology and Applied Pharmacology 19:416.
- Donohue, J. M., A. J. Piluso, and J. R. Schreiber. 1966. Acrolein—a biocide for slime control in cooling water systems. Materials Protection 5:22–24.

- Egle, J. L., Jr., and P. M. Hudgins. 1974. Dose-dependent sympathomimetic and cardioinhibitory effects of acrolein and formaldehyde in the anesthetized rat. Toxicology and Applied Pharmacology 28:358–366.
- Ferguson, F. F., C. S. Richards, and J. R. Palmer. 1961. Control of Australorbis glabratus by acrolein in Puerto Rico. U.S. Department of Health Education and Welfare, Public Health Reports 76:461-468.
- Feron, V. J., and A. Kruysse. 1977. Effects of exposure to acrolein vapor in hamsters simultaneously treated with benzo[a] pyrene or diethylnitrosamine. Journal of Toxicology and Environmental Health 3:379–394.
- Feron, V. J., A. Kruysse, H. P. Til, and H. R. Immel. 1978. Repeated exposure to acrolein vapour: subacute studies in hamsters, rats and rabbits. Toxicology 9:47–57.
- Fischer, R. F. 1962. Polymers from acrolein. Pages 225–233 in C. W. Smith, editor. Acrolein. John Wiley, New York.
- Folmar, L. C. 1976. Overt avoidance reaction of rainbow trout fry to nine herbicides. Bulletin of Environmental Contamination and Toxicology 15:509-514.
- Folmar, L. C. 1977. Acrolein, dalapon, dichlobenil, diquat, and endothal: bibliography of toxicity to aquatic organisms. U.S. Fish and Wildlife Service Technical Paper 88. 16 pp.
- Folmar, L. C. 1978. Avoidance chamber responses of mayfly nymphs exposed to eight herbicides. Bulletin of Environmental Contamination and Toxicology 19:312-318.
- Folmar, L. C. 1980. Effects of short-term field applications of acrolein and 2,4-D (DMA) on flavor of the flesh of rainbow trout. Bulletin of Environmental Contamination and Toxicology 24:217-224.
- Fritz-Sheridan, R. P. 1982. Impact of the herbicide Magnacide-H (2-propenal) on algae. Bulletin of Environmental Contamination and Toxicology 28:245–249.
- Fujino, M., T. Arima, Y. Sato, and H. Tako. 1985. Suppression of action of some excitation-contraction (E-C) uncoupling agents and mechanism in E-C coupling in single skeletal muscle fibers of frog, R. japonica. Journal of the Physiological Society of Japan 47:519.
- Green, M. A., and J. L. Egle, Jr. 1983. The effects of acetaldehyde and acrolein on blood pressure in guanethidine-pretreated hypertensive rats. Toxicology and Applied Pharmacology 69:29-36.
- Heck, H. d'A., M. Casanova, M. J. McNulty, and C. W.
 Lam. 1986. Mechanisms of nasal toxicity induced by formaldehyde and acrolein. Pages 235-247 in C. S.
 Barrow, editor. Toxicology of the nasal passages.
 Hemisphere Publishing Company, New York.
- Hill, R. D. 1960. The use of acrolein; acryladehyde; 2-propenal in the treatment of submerged weeds in farm ponds. Ohio Agricultural Experiment Station, Mimeographed 3 pp.
- Holcombe, G. W., G. L. Phipps, A. H. Sulaiman, and A. D. Hoffman. 1987. Simultaneous multiple species testing: acute toxicity of 13 chemicals to 12 diverse freshwater amphibian, fish, and invertebrate families.

- Archives of Environmental Contamination and Toxicology 16:697-710.
- Hudson, R. H., R. K. Tucker, and M. A. Haegele. 1984.
 Handbook of toxicity of pesticides to wildlife. U.S.
 Fish and Wildlife Service Resource Publication 153.
 90 pp.
- Kankaanpaa, J., E. Elovaara, K. Hemminki, and H. Vainio. 1979. Embryotoxicity of acrolein, acrylonitrile, and acrylamide in developing chick embryos. Toxicology Letters 4:93-96.
- Kaplan, H. L. 1987. Effects of irritant gases on avoidance/escape performance and respiratory response of the baboon. Toxicology 47:165–179.
- Kissel, C. L., J. L. Brady, A. M. Guerra, M. J. Meshishnek, B. A. Rockie, and F. F. Caserio, Jr. 1981.
 Monitoring acrolein in naturally occurring systems.
 Pages 102–116 in J. L. Johnson, J. R. Stanford, C. C. Wright, and A. G. Ostroff, editors. Water for subsurface injection: proceedings of the second symposium.
 ASTM STP 735, American Society for Testing and Materials, Philadelphia, Pennsylvania.
- Kobbia, I. A. 1982. Response of phytoplankton populations in some Egyptian irrigation drains to the aquatic weed herbicide "acrolein". Egyptian Journal of Botany 25:41-67.
- Korhonen, A., K. Hemminki, and H. Vainio. 1983. Embryotoxic effects of acrolein, methacrylates, guanidines and resorcinol on three day chicken embryos. Acta Pharmacologica et Toxicologica 52:95–99.
- Leach, C. L., N. S. Hatoum, H. V. Ratajczak, and J. M. Gerhart. 1987. The pathologic and immunologic effects of inhaled acrolein in rats. Toxicology Letters 39:189–198.
- Leikauf, G. D., L. M. Leming, J. R. O'Donnell, and C. A. Doupnik. 1989. Bronchial responsiveness and inflammation in guinea pigs exposed to acrolein. Journal of Applied Physiology 66:171-178.
- Lijinsky, W. 1988. Chronic studies in rodents of vinyl acetate and compounds related to acrolein. Annals of the New York Academy of Sciences 534:246–254.
- Lijinsky, W., and M. D. Reuber. 1987. Chronic carcinogenesis studies of acrolein and related compounds. Toxicology and Industrial Health 3:337-345.
- Lyon, J. P., L. J. Jenkins, Jr., R. A. Jones, R. A. Coon, and J. Siegel. 1970. Repeated and continuous exposure of laboratory animals to acrolein. Toxicology and Applied Pharmacology 17:726-732.
- Macko, V., J. A. A. Renwick, and J. F. Rissler. 1978. Acrolein induces differentiation of infection structures in the wheat stem rust fungus. Science 199:442-443.
- Marano, F., and S. Puiseux-Dao. 1982. Acrolein and cell cycle. Toxicology Letters 14:143–149.
- Mayer, F. L. 1987. Acute toxicity handbook of chemicals to estuarine organisms. U.S. Environmental Protection Agency, Report EPA/600/8–87/017. 274 pp.
- McKim, J. M., P. K. Schmeider, G. J. Niemi, R. W. Carlson, and T. R. Henry. 1987. Use of respiratory-

- cardiovascular responses of rainbow trout (Salmo gairdneri) in identifying acute toxicity syndromes in fish: part 2. Malathion, carbaryl, acrolein and benzaldehyde. Environmental Toxicology and Chemistry 6:313-328.
- Mitchell, D. Y., and D. R. Petersen. 1988. Inhibition of rat liver aldehyde dehydrogenases by acrolein. Drug Metabolism and Disposition 16:37-42.
- National Institute for Occupational Safety and Health (NIOSH). 1990.
- NIOSH pocket guide to chemical hazards, U.S. Department of Health and Human Services, Public Health Service, Centers for Disease Control:32-33.
- National Research Council (NRC) 1977. Acrolein. Pages 553-556 in Drinking water and health. National Academy of Sciences, Washington, D.C.
- Nishikawa, I., and T. Hayakawa. 1986. Bromination and gas chromatographic determination of micro amounts of acrolein in rain water. Journal of Chromatography 551:566-570.
- O'Loughlin, E. M., and K. H. Bowmer. 1975. Dilution and decay of aquatic herbicides in flowing channels. Journal of Hydrology 26:217-235.
- Parent, R. A., H. E. Caravello, and J. E. Long. 1992. Two-year toxicity and carcinogenicity study of acrolein in rats. Journal of Applied Toxicology 12:131-139
- Parker, L. M., D. J. Lauren, J. B. Cooke, and D. E. Hinton. 1990. Metabolism of endogenous and xenobiotic aldehydes by rainbow trout (Oncorhynchus mykiss) liver fractions. Aquatic Toxicology 18:1-12.
- Reinert, K. H., and J. H. Rodgers. 1987. Fate and persistence of aquatic herbicides. Reviews of Environmental Contamination and Toxicology 98:61-98.
- Rijstenbil, J. W., and G. C. van Galen. 1981. Chemical control of mussel settlement in a cooling water system using acrolein. Environmental Pollution 25A:187-195.
- Sanduja, R., G. A. Ansari, and P. J. Boor. 1989. 3-hydroxypropylmercapturic acid: a biologic marker of exposure to allylic and related compounds. Journal of Applied Toxicology 9:235-238.

- Sierra, L. M., A. R. Barros, M. Garcia, J. A. Ferreiro, and M. A. Comendador. 1991. Acrolein genotoxicity in Drosophila melanogaster. I. Somatic and germinal mutagenesis under proficient repair conditions. Mutation Research 260:247-256.
- Sine, C., editor. 1991. Farm chemicals handbook '91. Meister Publishing Company, Willoughby, Ohio.
- Slott, V. L., and B. F. Hales. 1986. The embryolethality and teratogenicity of acrolein in cultured rat embryos. Teratology 34:155-163.
- Slott, V. L., and B. F. Hales. 1987. Protection of rat embryos in culture against the embryotoxicity of acrolein using exogenous glutathione. Biochemical Pharmacology 36:2187-2194.
- Smith, C. W., editor. 1962. Acrolein. John Wiley, New York. 273 pp.
- Srivastava, S. C., R. K. Upreti, and A. M. Kidwai. 1992. Action of acrolein on rat liver membrane proteins and enzymes. Bulletin of Environmental Contamination and Toxicology 49:98-104.
- Starzecka, A. 1975. The influence of acrolein and hydrocryle on the development dynamics of aquatic bacteria. Acta Hydrobiologia 17:391-403.
- Susten, A. S., and M. J. Breitenstein. 1990. Failure of acrolein to produce sensitization in the guinea pig maximization test. Contact Dermatitis 22:299-300.
- Toraason, M., M. E. Luken, M. Brietenstein, J. A. Krueger, and R. E. Biagini. 1989. Comparative toxicity of allylamine and acrolein in cultured myocytes and fibroblasts from neonatal rat heart. Toxicology 56:107-117.
- U.S. Environmental Protection Agency (EPA). 1980. Ambient water quality criteria for acrolein. Report EPA 440/5-80-016. 94 pp.
- Warholm, M., B. Holmberg, J. Hogberg, T. Kronevi, and A. Gotharson. 1984. The acute effects of single and repeated injections of acrolein and other aldehydes. International Journal on Tissue Reactions 6:61-70.
- Weber-Tschopp, A., T. Fischer, R. Gierer, and E. Grandjean. 1977. Experimentelle reizwirkungen von akrolein auf den menschen. International Archives of Occupational and Environmental Health 40:117-130.

| | Publication | Biological Report |
|---------------------|---------------|-------------------|
| Subject | date | number |
| Mirex | March 1985 | 85(1.1) |
| Cadmium | July 1985 | 85(1.2) |
| Carbofuran | August 1985 | 85(1.3) |
| Toxaphene | August 1985 | 85(1.4) |
| Selenium | October 1985 | 85(1.5) |
| Chromium | January 1986 | 85(1.6) |
| Polychlorinated | • | 33(2.0) |
| Biphenyls | April 1986 | 85(1.7) |
| Dioxins | May 1986 | 85(1.8) |
| Diazinon | August 1986 | 85(1.9) |
| Mercury | April 1987 | 85(1.10) |
| Polycyclic Aromatic | • | 33(1.10) |
| Hydrocarbons | May 1987 | 85(1.11) |
| Arsenic | January 1988 | 85(1.12) |
| Chlorpyrifos | March 1988 | 85(1.13) |
| Lead | April 1988 | 85(1.14) |
| Tin | January 1989 | 85(1.15) |
| Index to Species | February 1989 | 85(1.16) |
| Pentachlorophenol | April 1989 | 85(1.17) |
| Atrazine | May 1989 | 85(1.18) |
| Molybdenum | August 1989 | 85(1.19) |
| Boron | April 1990 | 85(1.20) |
| Chlordane | July 1990 | 85(1.21) |
| Paraquat | August 1990 | 85(1.22) |
| Cyanide | December 1991 | 85(1.23) |
| Fenvalerate | May 1992 | 2 |
| Diflubenzuron | June 1992 | 4 |
| Zinc | April 1993 | 10 |
| Famphur | February 1994 | 20 |
| Acrolein | June 1994 | 23 |

^a Copies of individual reviews, if available, may be obtained gratis from the Publications Unit, U.S. Fish and Wildlife Service, 1849 C Street, N.W., Mail Stop 130—ARLSQ, Washington, DC 20240, or may be purchased from the National Technical Information Service (NTIS), 5285 Port Royal Road, Springfield, Virginia 22161.

U.S. Department of the Interior National Biological Survey

As the Nation's principal conservation agency, the Department of the Interior has responsibility for most of our nationally owned public lands and natural resources. This responsibility includes fostering the sound use of our lands and water resources; protecting our fish, wildlife, and biological diversity; preserving the environmental and cultural values of our national parks and historical places; and providing for the enjoyment of life through outdoor recreation. The Department assesses our energy and mineral resources and works to ensure that their development is in the best interests of all our people by encouraging stewardship and citizen participation in their care. The Department also has a major responsibility for American Indian reservation communities and for people who live in island territories under U.S. administration.